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**South Central
Connecticut
Regional Water
Authority**

2008 BENTHIC BIOLOGICAL ASSESSMENT OF THE LOWER MILL RIVER, HAMDEN/NEW HAVEN, CT



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Prepared By:



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INTRODUCTION

The purpose of this study is to provide aquatic resource management information concerning resumption of water withdrawals from Lake Whitney and possible alterations to stream flows in the downstream Mill River. The study provides quantitative and qualitative information about general habitat characteristics and benthic macroinvertebrate community structure at five locations along the lower Mill River in Hamden and New Haven, CT downstream of Lake Whitney. This study summarizes survey results from 2008. In April 2005 the new water treatment facility which draws water from Lake Whitney went online, and this study represents the third year of post-operational data collection. The water treatment facility was operating mostly in a testing mode in 2005, and withdrawals were generally near the low end of the expected range, averaging 16 percent of the maximum allowed withdrawal. 2006 operations consisted of higher but still very moderate withdrawal rates, averaging 31% of the maximum allowed withdrawal. The most significant flow alteration in 2006 occurred as a result of lowering the water level of Lake Whitney for a water supply construction project. This resulted in downstream flows exceeding natural inflow during the drawdown period, followed by a period of minimum downstream releases after the project while the reservoir refilled. 2007 operations consisted of lower withdrawal rates, averaging only 28% of the maximum allowed withdrawal. In 2007, Lake Whitney was drawn down for dam inspection and maintenance on two occasions (June and October), for a total of 12 days. Withdrawals in the first month of 2008 were up to 85% of maximum allowed withdrawal due to low reservoir storage system-wide but were significantly cut back as record rainfall in 2008 (highest in 97 year period of record at Whitney rain gauge) restored system storage to significantly above average levels. From late October 2008 through the end of the year, treatment plant operation was reduced to one day per week. Overall, average withdrawals in 2008 were 30% of the maximum allowed withdrawal. It is intended that a review of all data collected in 2005, 2006, 2007 and 2008, as well as future operational years will be conducted to evaluate any potential impacts to Mill River from the water withdrawal in Lake Whitney. Ultimately, post-operation data will be compared to pre-operation data collected in 1998 and 2000 to 2004. This investigation facilitates that analysis, but focuses on extending the post-operational database. ENSR Corporation, a subsidiary of AECOM, has recently adopted the name "AECOM Environment," but the staff doing assessment work in relation to this project remain the same.

METHODS

General methods were consistent with previous years, beginning in 2000. Samples were collected on June 6 and August 19, 2008, at the peak of the tidal outflow (low tide). Sampling locations (Figure 1) were the same as previous years, except in 2006 station 5 was eliminated from the monitoring program due to the tidal influences and to focus on more detailed chironomid analysis at more relevant upstream stations. Sampling stations were longitudinal stretches, ranging from 85 to 300 ft in length (~25-90 m). Each sampling station was characterized for general habitat and instream water quality at representative sites. A single sample per site was used to determine water quality features on the day of sampling. Flow values were daily means from SCCRWA flow records from the Whitney Dam.

Aquatic habitat was evaluated in a qualitative to semi-quantitative way. This was a modified version of the USEPA Rapid Bioassessment Protocol (Physical Characterization / Water Quality Assessment) (Barbour et al. 1999). Aquatic habitat characterization included features such as surrounding land use, canopy cover, flow, and substrate composition for each sampling station. Water quality was assessed in a quantitative way with in situ determinations of water temperature, dissolved oxygen content, conductivity, turbidity, and pH at each sampling station.

Timed (two minutes) D-frame dip-net sampling was used to collect macroinvertebrates. This method is commonly used as a multi-habitat rapid bioassessment technique (Barbour et al. 1999). Riffle habitats were sampled at stations where riffle habitat is available, although at higher flows some of these areas could be characterized as run habitats. Macroinvertebrates were captured in the net by dislodging the substrate up to 1 ft (0.3 m) upstream of the dip-net. Two subsamples per sampling station were collected. Each subsample consisted of a two-minute collection, itself comprised of four 30-second collection efforts at four nearby locations within the site. Subsamples were preserved in 70% ethanol for laboratory analysis. Macroinvertebrates were sorted, identified to the lowest practical taxonomic level, and counted. Samples were collected during the period of low tide on both sampling dates.

After a 2005 test run with a subset of the total sample collection, Chironomidae samples were identified in 2006, 2007 and 2008 to the lowest practical taxonomic division, typically the genus or species level, to further facilitate water quality analysis. Although the main focus of this monitoring program is on the impacts of changing flows, flow can affect water quality, and pollution tolerance of individual species varies within the Chironomidae family.

The two macroinvertebrate subsamples were analyzed separately, but combined into a single sample per station for data analysis. Variability among subsamples was evident, as is expected for such samples, but was not striking. Numerical analysis included relative abundance and dominance patterns on taxonomic and feeding group bases, species richness and diversity. Species richness was expressed as number of taxa (S). Species diversity quantifies the degree of dominance (or lack thereof) of taxa within a community; it measures the distribution of

individuals among taxa present. When one or a few taxa dominate a community, diversity is low. The Hilsenhoff Biotic Index (HBI), based on a quality value of 0-10 assigned to each taxon multiplied by the abundance of each corresponding taxon and divided by the total number of individuals was calculated for each station. Modified HBI calculations were completed for all data collected from 2000-2008. The index was modified to include non-arthropod species (Mandeville 2002).

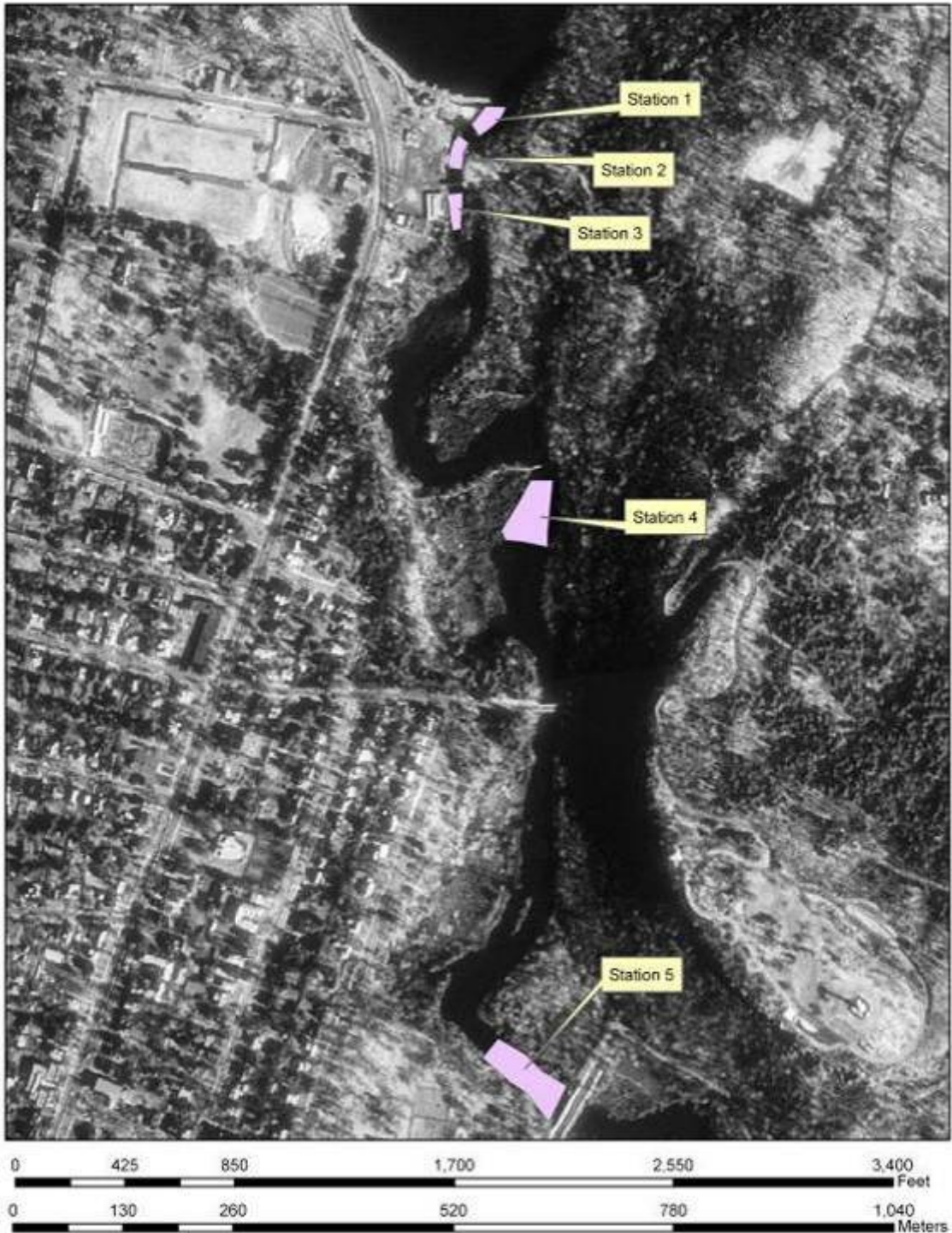


Figure 1. Locations of the five established sampling stations along the Lower Mill River in Hamden (stations 1-4) and New Haven (station 5). Station 5 was eliminated as an active biological sampling station in 2006.

RESULTS

Habitat Characterization

Predominant land use (forest and residential) and sources of pollution (storm pipes discharging at several locations between stations 2 and 4) were the same in 2008 as in all previous surveys (Table 1). Sources of pollution to the lower Mill River include a number of combined sewer overflows (CSOs), the closest to the study area being located at East Rock Road between Stations 4 and 5. CSOs can have strong but intermittent water quality impacts in the tidal areas of the river. Canopy cover reached its maximum at station 3 and its minimum at station 1. Major shore or bank erosion was not observed.

Flow is estimated by the SCCRWA using automated lake level measurements at the Lake Whitney spillway. Flows on the day of the survey are not necessarily an indication of antecedent conditions, however, and SCCRWA flow records were consulted to categorize the hydrological conditions for two and a half months before each sampling (CH2MHILL 2009). Based on factors such as tidal influence and watershed hydrologic characteristics, a wide range of flow conditions might be anticipated at any given time within the study area. Tidal influences are apparent at stations 3 and 4 as water level fluctuations. Variation in flow from Lake Whitney is the more dominant current influence at stations 1 and 2. Under low flow conditions, salinity influences at station 4 are measurable (CH2MHILL 2008). However, while water level changes with tide are evident at station 3, saltwater does not intrude this far upstream. In 2008 the average daily spring flow in the 10-week period preceding the June 6 sampling (94 mgd) was larger than the average daily summer flow preceding the August 19 sampling (43 mgd) (Table 2), as expected. Flows in the spring of 2007 were higher than any previously observed average daily 10 week flow preceding sampling since the inception of the study program. April of 2007 was the 2nd wettest April in the 96 year period of record of rainfall measurements at Lake Whitney. Spring flow values increased each year between 2005 and 2007, but decreased in 2008. Flows during spring 2008 were nearly identical to the 2004 flows. Despite elevated rainfall during the summer resulting in the 2nd wettest summer in the 97-year rain gauge record at Lake Whitney, the 10-week average flow for the period before the August 19th sampling event fell within the range of values recorded from 2005-2007. High rainfall in summer of 2008 kept minimum lake levels well above spillway elevation, with the lowest estimated average daily flow in summer 2008 being 17 mgd. This avoided the need to artificially release water from Lake Whitney.

The abundance and distribution of aquatic vegetation was similar to pre-operational years. The amount of filamentous algae and rooted aquatic plants varied among sampling locations in 2008 and is likely a function of varied flow. In 2008, the abundance of aquatic macrophytes as percent cover at each station was similar at all stations. Unlike 2007 (CH2MHILL 2008), station 4 was not influenced by saltwater intrusion in 2008 due to high precipitation and freshwater flows throughout the year (J. Hudak, personal communication).

Average stream depth and width were similar to previous years. Stream width and depth were on the higher end of stream width and depth in June, and on the lower end in August, due to variations in flow on the date of sampling. Tide influenced stream depth at Station 4. However, as sampling at station 4 was conducted under low tide conditions, observed fluctuations were minor in comparison with possible changes over the tidal cycle.

Inorganic substrates were generally coarser at the upstream sites (Stations 1 and 2) and progressively decreased in mean particle size in the downstream direction (Table 1). Fine-grained substrate such as silt was observed only at the most downstream station (i.e., Station 4). Data from previous years suggest particle transport is occurring during large storm events, but the amount of transport has not been examined.

Detritus (e.g., logs, wood, leaf litter) was present at relatively low levels, indicating periodic flushing as would be expected in this large watershed. Most stations had similar percentages of detritus. Station 4 had the greatest amount of detritus, but the relative amount was minimal in comparison with inorganic substrates. However, general amounts of detritus, both fine and coarse, appeared to be sufficient to support abundant populations of macroinvertebrates at all stations.

Vegetation levels in 2008 were similar to those in previous survey years, with increased macrophyte growth at stations 3 and 4, compared to stations 1 and 2. Our experiences from previous years is that species tolerant of high flow such as attached moss and filamentous green algae (Chlorophyta: Chlorophyceae) comprised the majority of the vegetation at the upstream stations (1 and 2), but presence of rooted macrophytes (mostly narrow-leaved pondweeds) was noted in the upstream area during some samplings. Filamentous algal abundance at stations 1 and 2 decreased between June and August samplings, perhaps in response to decreasing flows. Stations 1 and 2 experienced an overall decrease in macrophyte abundance between June and August related to a narrow river channel under decreased flows, a pattern observed nearly every year.

Waterlilies (*Nymphaea* sp., a freshwater species that prefers slow-flowing to lentic waters) were observed at the downstream stations. All the taxa of vascular plants encountered in the lower Mill River in 2008 were common taxa, tolerant of conditions such as low light, high nutrients, and salinity gradients (Crow and Hellquist 1980). Total plant coverage at the sites was within the typical ranges observed for temperate lotic systems (Allan 1995), and species present between years are similar.

In general, habitat structure was suitable for macroinvertebrates at all stations in 2008. Substrate structural complexity (i.e., spatial heterogeneity) provides a diverse habitat for invertebrates, creating “niches” dominated by different food resources and hence varied invertebrate species, and/or providing crevices that protect invertebrates from predation or

complete dislodgement by strong currents (Hixon & Menge 1991; Allan 1995). Macrophytes also contribute to increased spatial heterogeneity by providing a substrate rich in food resources (epiphytic algae and detritus covering the plants) (Diehl & Kornijów 1998). Physical substrate (cobble and gravel substrate) and/or macrophyte cover was sufficient to potentially support a rich and diverse macroinvertebrate community at all stations. As seen in previous years, habitat quality at station 4 was not as high as at stations 1-3 in 2008.

Selected water quality parameters were assessed in 2008 during both sampling events (Table 2). Assessed water quality in 2008 was slightly different than previous years for some parameters. The pH of most samples was slightly basic to basic (Table 2). All stations had August pH values higher than any value measured previously, except for Station 1. Values for pH in 2008 increased between June and August at all stations, but remained well within the life compatible 4.5 – 9.5 range for most aquatic biota (Wetzel 2001b). Dissolved oxygen levels were above the Connecticut Water Quality Standard of 5 mg/L at all stations during the sampling in June and August, considered adequate to support aquatic life. AECOM did not observe dissolved oxygen levels below 5.0 mg/L in 2008. A detailed study of dissolved oxygen in the study area indicated that individual readings were below 5.0 mg/L, but the overall average was over 5.0 mg/L (CH2MHILL 2009).

Salinity levels at Station 4 were about 0.1 ppt in both June and August and lower than measured salinities from 2005, 2006 and 2007. Water temperature in 2008 was within the range from previous years. Water temperature in August was higher than in June, which is typical.

Specific conductivity was comparable between stations during June and August. Saltwater influence at station 4 has been responsible for increased conductivity in previous years, but low salinity values in 2008 did not have a significant impact on conductivity. There is evidence of saltwater intrusion at lower flows during dry summers, extending just upstream of Station 4 (CH2MHill 2001).

Turbidity varied among stations and dates to some degree, but was generally low to moderate at the time of sampling. Very high turbidity is known from the Mill River system upstream of Lake Whitney, but the lake acts as a detention basin and minimizes downstream transport of particles much of the time. During the August 2008 sampling event, water at stations 1-3 had a green tint, but this did not translate into high turbidity levels.

Table 1. - Lower Mill River habitat characterization. Data are for the June and August sampling events in 2008.

Parameters	Stn 1		Stn 2		Stn 3		Stn 4		Stn 5	
	Jun 6	Aug 19	Jun 6	Aug 19	Jun 6	Aug 19	Jun 6	Aug 19	Jun 6	Aug 19
Length of Segment	85 ft (26 m)		150 ft (46 m)		300 ft (91 m)		300 ft (91 m)		300 ft (91 m)	
Watershed/Bank Features										
predominant surrounding land use	forest/residential		forest/residential		forest/residential		forest/residential		forest/residential	
canopy cover	open		some shade (<40%)		mod. Shade (30-80%)		some shade (<40%)		some shade (<40%)	
dominant riparian vegetation	shrubs		shrubs		trees		trees/shrubs		trees	
bank stability ⁽¹⁾	stable		stable		stable		stable		stable	
other notable features	near dam		near dam		downstream of dam		tidal influence		tidal influence	
In-stream Features										
<u>general habitat type (%)</u>										
riffle	100	100	85	90	90	100	-	-	-	-
run	-	-	15	10	10	-	85	40	-	-
pool	-	-	-	-	-	-	15	60	-	-
estimated stream width (ft):	80	40	75	30	100	75	130	85	-	-
estimated stream depth (ft):										
riffle	1.5	0.5	1.5	0.75	0.5	0.3	-	-	-	-
run	-	-	1.0	1.0	0.8	-	3.0	2.5	-	-
pool	-	-	-	-	-	-	3.5	3.5	-	-
<u>inorganic substrate composition⁽²⁾</u>										
bedrock	-	-	-	-	-	-	-	-	-	-
boulder (>256 mm)	10	10	10	10	0	5	5	5	-	-
cobble (64-256 mm)	75	75	60	60	20	20	10	10	-	-
gravel (2-64 mm)	15	15	30	20	65	65	30	30	-	-
sand (0.06-2 mm)	-	-	-	10	15	15	30	30	-	-
silt (0.004-0.006 mm)	-	-	-	-	-	-	25	25	-	-
clay (<0.004 mm)	-	-	-	-	-	-	-	-	-	-
<u>organic substrate composition⁽²⁾</u>										
detritus ⁽³⁾	0	5	5	10	15	15	25	20	-	-
aquatic macrophytes (total)	40	30	75	55	40	40	45	40	-	-
filamentous algae	100	100	40	60	75	80	20	15	-	-
water lilies (<i>Nymphaea</i> , <i>Nuphar</i>)	-	-	-	-	-	5	45	40	-	-
pondweeds (<i>Potamogeton spp</i>) ⁽⁴⁾	-	-	60	40	15	15	20	35	-	-
moss	-	-				-		-		
waterweed (<i>Elodea canadensis</i>)	-	-			10	-	15	5	-	-
tidal influence	No	No	No	No	No	No	Yes	Yes	-	-

(1) stable = minimal evidence of erosion or bank failure

(2) percent coverage

(3) logs, wood, coarse particulate organic matter

(4) narrow-leaved species.

Table 2. Water quality ranges and flows at the sampling locations in 2008. Pre-operation data is also presented as a range of values over all pre-operation years.

Parameter	Station 1					
	Pre-operation Range		Jun 6 2008		Aug 19 2008	
	Jun	Aug				
water temperature (°C)	17.9-23.2	19.8-26.7	19.8		23.6	
dissolved oxygen (mg/L)	8.3-9.7	5.7-9.4	8.8		8.6	
dissolved oxygen (% saturation)	99-112	71-108	96.1		101.2	
specific conductivity (µS/cm)	189-282	194-270	242		210	
turbidity (NTU)	1.0-3.2	1.6-5.6	1.2		3.1	
pH (SU)	7.2-8.5	6.8-8.4	7.8		8.6	
Flow (mgd) (Average daily flow over prior 10 weeks)	88-140	42-97	94		43	
Parameter	Station 2					
	Pre-operation Range		Jun 6 2008		Aug 19 2008	
	Jun	Aug				
water temperature (°C)	17.7-23.2	19.7-26.4	19.8		23.7	
dissolved oxygen (mg/L)	8.0-10.4	7.3-9.0	8.7		8.4	
dissolved oxygen (% saturation)	94-120	86-111	94.8		98.5	
specific conductivity (µS/cm)	190-284	192-268	242		212	
turbidity (NTU)	1.0-7.9	1.2-7.8	1.4		3.3	
pH (SU)	7.2-8.5	7.6-8.8	7.8		8.8	
Flow (mgd) (Average daily flow over prior 10 weeks)	88-140	42-97	94		43	
Parameter	Station 3					
	Pre-operation Range		Jun 6 2008		Aug 19 2008	
	Jun	Aug				
water temperature (°C)	17.6-23.3	19.7-26.7	19.7		23.5	
dissolved oxygen (mg/L)	7.9-10.2	5.9-9.3	8.7		9.7	
dissolved oxygen (% saturation)	93-117	73-109	95.0		114.0	
specific conductivity (µS/cm)	189-290	194-265	242		213	
turbidity (NTU)	1.2-3.8	1.6-4.8	1.7		3.6	
pH (SU)	7.2-8.6	7.6-8.2	7.8		8.8	
Flow (mgd) (Average daily flow over prior 10 weeks)	88-140	42-97	94		43	
Parameter	Station 4					
	Pre-operation Range		Jun 6 2008		Aug 6 2008	
	Jun	Aug	Surface	Bottom	Surface	Bottom
water temperature (°C)	17.8-23.5	19.7-30.2	19.5	19.6	22.9	22.7
dissolved oxygen (mg/L)	7.9-11.8	6.1-8.9	8.5	8.5	6.9	6.2
dissolved oxygen (% saturation)	92-134	72-117	92.7	92.8	80.0	71.5
specific conductivity (µS/cm)	189-290	194-7013	242	242	218	220
turbidity (NTU)	1.2-4.6	1.9-8.4	2.1	-	2.8	-
pH (SU)	7.3-8.8	7.2-8.3	7.8	7.8	8.4	8.4
Salinity (ppt)	-	-	0.12	0.12	0.11	0.11
Flow (mgd) (Average daily flow over prior 10 weeks)	88-140	42-97	94	94	43	43

Macroinvertebrates

This investigation focused on the invertebrate community as an indicator of conditions downstream of Lake Whitney. Invertebrates have long been used as indicators of environmental quality, and will reflect water quantity effects to the extent that water quantity affects water quality (e.g., dilution, runoff). In the extremes, water quantity can also affect invertebrates by altering the substrate (scouring or drying/oxidation), through dislodgment of biota with downstream transport, and through reduced available habitat under dry conditions. Most effects of water quantity are indirect, however, necessitating a considerable database to allow an analysis that accounts for other potentially influential factors. An initial survey of the Mill River downstream of Lake Whitney was conducted in 1998, from which it was determined that invertebrates might provide suitable indication of the impact of changing flow as a consequence of the re-activation of Lake Whitney as a water supply.

2008 raw data for benthic macroinvertebrates have been analyzed in several ways relevant to questions of flow impacts. Total benthic macroinvertebrate abundance in 2008 (Figure 3) varied within and among stations. The obvious conclusion for previous years is that invertebrates are more abundant at stations 1-3 than at station 4. In 2008, this pattern of decreased abundance in the downstream direction was observed in June, but was not observed in August. In August 2008, stations 1 and 2 had lower abundance than station 4. There are both physical and chemical habitat changes between stations 3 and 4 that are more likely to be responsible for this difference than any variation in flow. Although not observed on the days of sampling, the primary influence for decreased abundance is likely tidal, with slower water velocities, changing direction of flow, and oscillating salinity at station 4.

In 2008 there was a decrease in invertebrate abundance at all stations in June and August compared to 2007, except for station 3 in August where abundance increased. Invertebrate abundance in 2008 increased between June and August for stations 2, 3 and 4, but decreased slightly at station 1. In 2005 we witnessed the largest numbers of invertebrates since the inception of the study program at stations 1 and 2 (Figures 3 and 4), and high numbers of invertebrates were observed in 2006 and 2007. In 2008, abundance levels were within the range of values observed previously for all stations (Figure 4).

Taxonomically, the assemblage of invertebrates in the Mill River downstream of Lake Whitney exhibits variable richness (Figure 5), with between 7 and 13 taxa identified at each station for June 2008 and between 5 and 10 taxa for August 2008. The findings in 2008 are comparable to previous years where the number of taxa present at each station varied between 6 and 28, although station 4 in 2008 had the lowest number of taxa ever observed since study inception. 2008 richness has a smaller range of values compared to the three previous years with the treatment facility online. Richness in 2005, 2006 and 2007 ranged between 6 and 17 taxa, 10 and 16 taxa, and 7 and 21 taxa respectively. This assessment excludes the detailed Chironomidae investigation: chironomids have only been identified below the family level at all stations beginning in 2006, and will be addressed separately.

A cumulative look at the abundance of invertebrates within the more common taxa encountered in 2005, 2006, 2007 and 2008 (Figure 6), indicates that the most common taxon (the caddisfly *Macrostemum*) is by far the most abundant. The next two most abundant taxa are the midges in the family Chironomidae, and an oligochaete worm, *Nais communis*. The 15 most abundant taxa are shown in Figure 6, with the next 10 most abundant lumped together and the remaining 32 taxa lumped into yet another category for graphic comparison. Adding a fourth year of post-operational data did not result in any changes to the top four most common taxa. The midges, *Nais communis*, and *Gammarus* sp. are still the 2nd, 3rd and 4th most abundant taxa after the addition of the 2008 data. In 2008, only 19 *Macrostemum* were collected, but extremely high abundances in previous years resulted in no change to the rank.

In general, the common taxa observed in any one year were also encountered in the other years, although rare taxa are encountered during sampling. In 2005, two new taxa were collected, *Donacia* (leaf beetles) and *Neophylax* (caddisfly). *Donacia* has not been observed since 2005, but *Neophylax* has been collected in 2006, 2007 and 2008. In 2007, 16 new taxa were collected in the Mill River, but most were present in low numbers. In 2008, members of the family Perlidae (stonefly) were collected for the first time. Individuals were collected in low numbers at stations 1 and 2 in June, and stations 1-3 in August. In previous years we found that less common taxa were not consistently observed over time or space. Rare taxa tend to be patchily distributed, and patchiness may be exacerbated by spatial habitat heterogeneity. Therefore, absence of such rare taxa in some samples or years may not mean that the taxa were not present in the lower Mill River system.

An alternative way to evaluate the macroinvertebrate data is to organize them by feeding groups. These groups have ecological meaning in terms of food resources and energy flow, and may be affected by flow insofar as flow affects food delivery from upstream, the growth of periphyton, and the accumulation of organic detritus. 2008 feeding group data varied between stations and among sampling dates (Figures 7-10). Stations 1-3 were dominated by collectors and shredders with some scrapers, while collectors and predators were most abundant at station 4. General patterns of feeding group abundance between post-operational years (2005-2007) appear similar, although slight shifts are present based on specific species occurrences.

Hilsenhoff Biotic Index values at each station were calculated and graphed against the 10-week average flows prior to sampling for each year (Figures 11-15). The graphs do not include the HBI values for the September 2004 sampling event due to the Lake Whitney drawdown for upgrades to the dam related to the new treatment facility. HBI values for 2008 are within the range of values observed previously. Values for all years ranged from 4.6-8.2 at Station 1, 3.7-7.0 at station 2, 4.7-7.2 at station 3, 5.5-9.0 at station 4 and 5.9-7.4 at station five (Table 3).

To assess the impacts of the water treatment facility on the invertebrate community in the Mill River, pre-operation and post-operation data were grouped separately and graphed against the

flows for each sampling occasion. Flow was graphed against taxonomic richness, total individuals, evenness and diversity (Figures 16-19). Diversity values are affected by the number of taxa present at each station, while evenness is a normalized measure of diversity that puts all values on a scale of zero (low) to one (high). Pre-operation and post-operation data is similar for taxonomic richness and diversity. Evenness between pre and post-operation data is similar, although post-operational data appears to be slightly higher. Total number of individuals varies between data sets. Four of the five highest numbers of individuals at any station or date have occurred since the water treatment facility went online. Although slight differences in the data may be suggested visually, no trend in flow impacts is apparent.

Lake Whitney Overflow

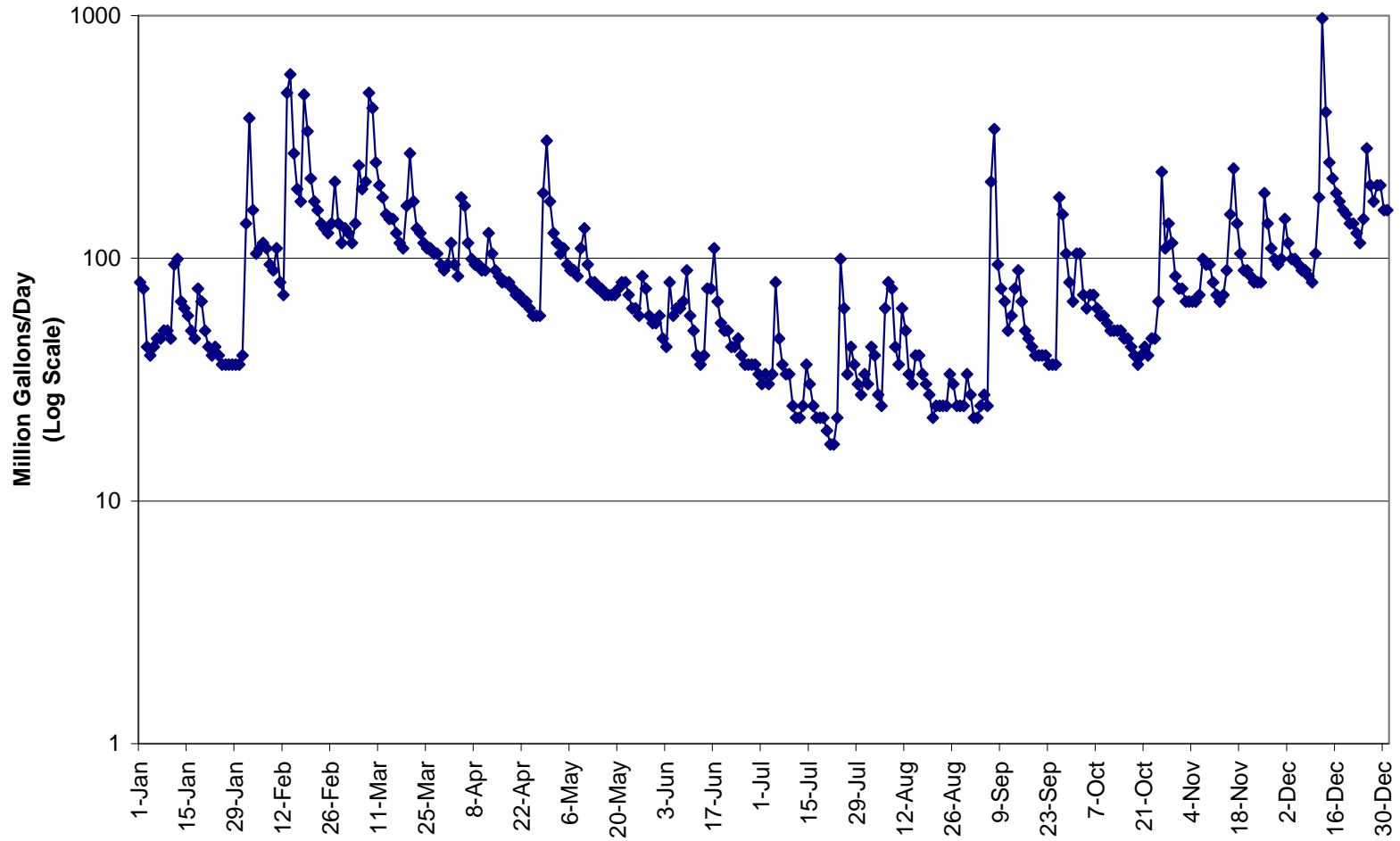


Figure 2. Mill River flows in 2008 measured at the Lake Whitney spillway.

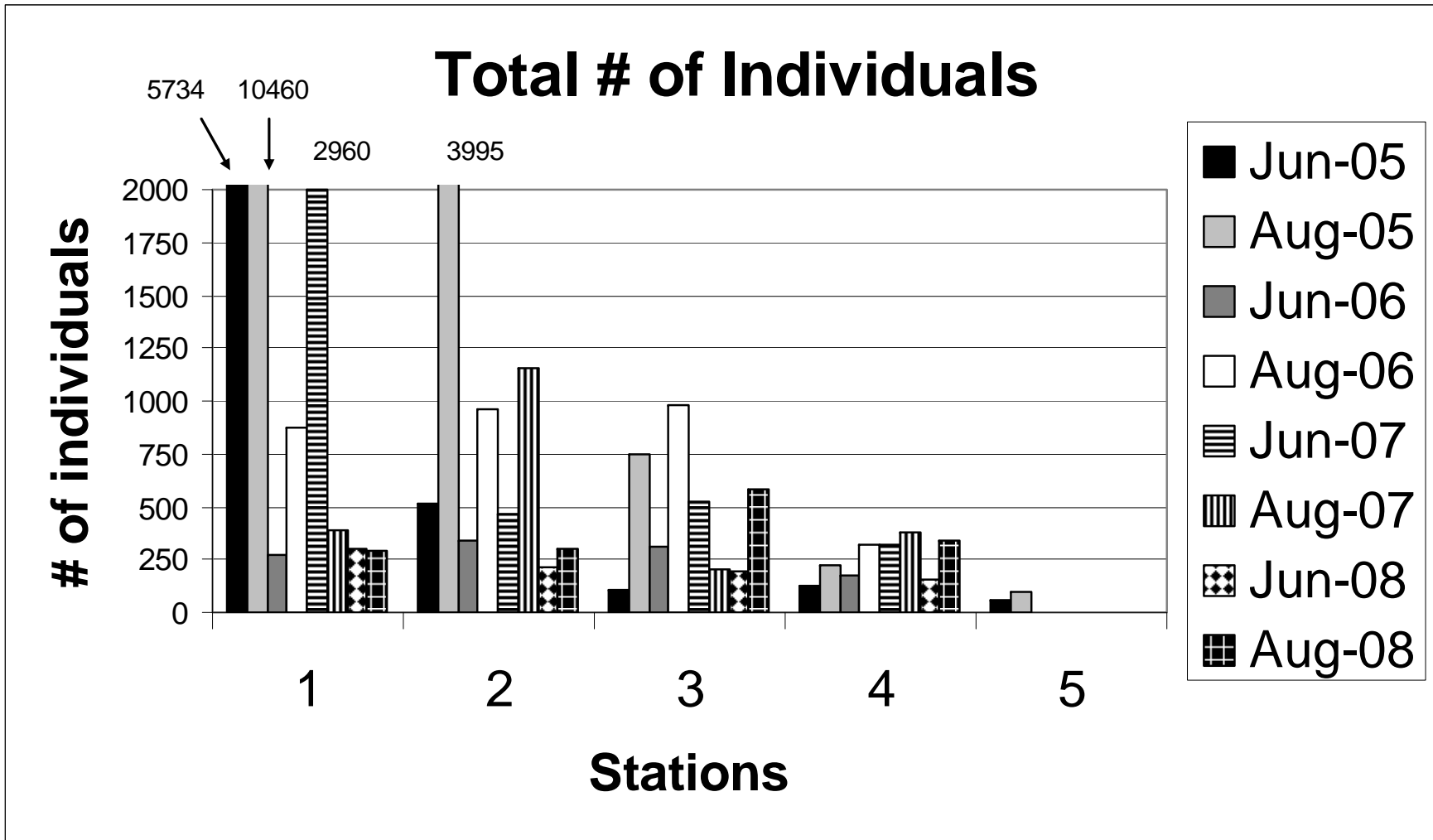


Figure 3. 2005, 2006, 2007 and 2008 benthic macroinvertebrate abundance over space and time in the Mill River, downstream of Lake Whitney. Macroinvertebrate abundance is based on two timed, two minute D frame net samples.

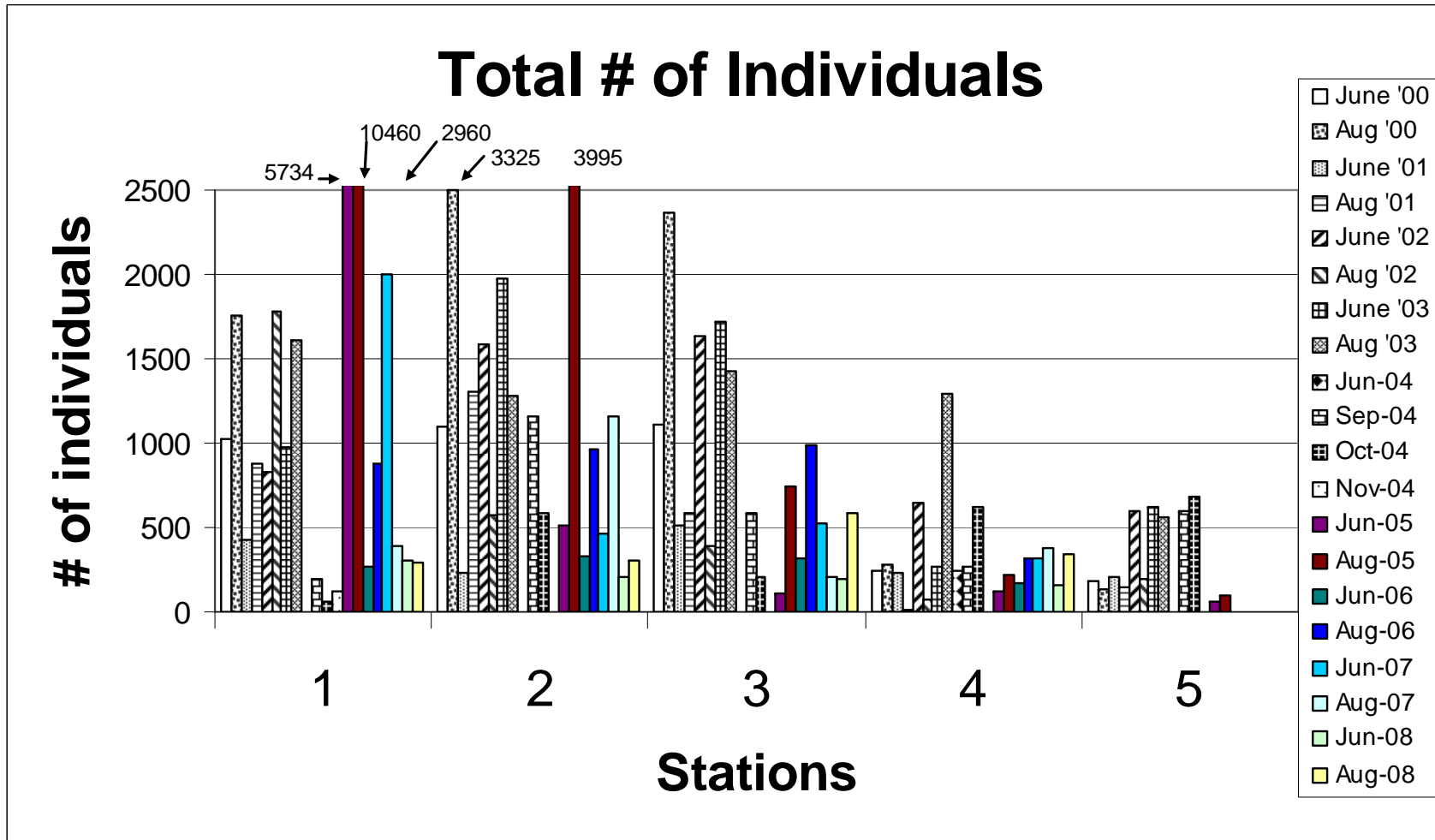


Figure 4. Total number of invertebrates over space and time in the Mill River, downstream of Lake Whitney for all years. These values are based on two timed, two minute D frame net samples.

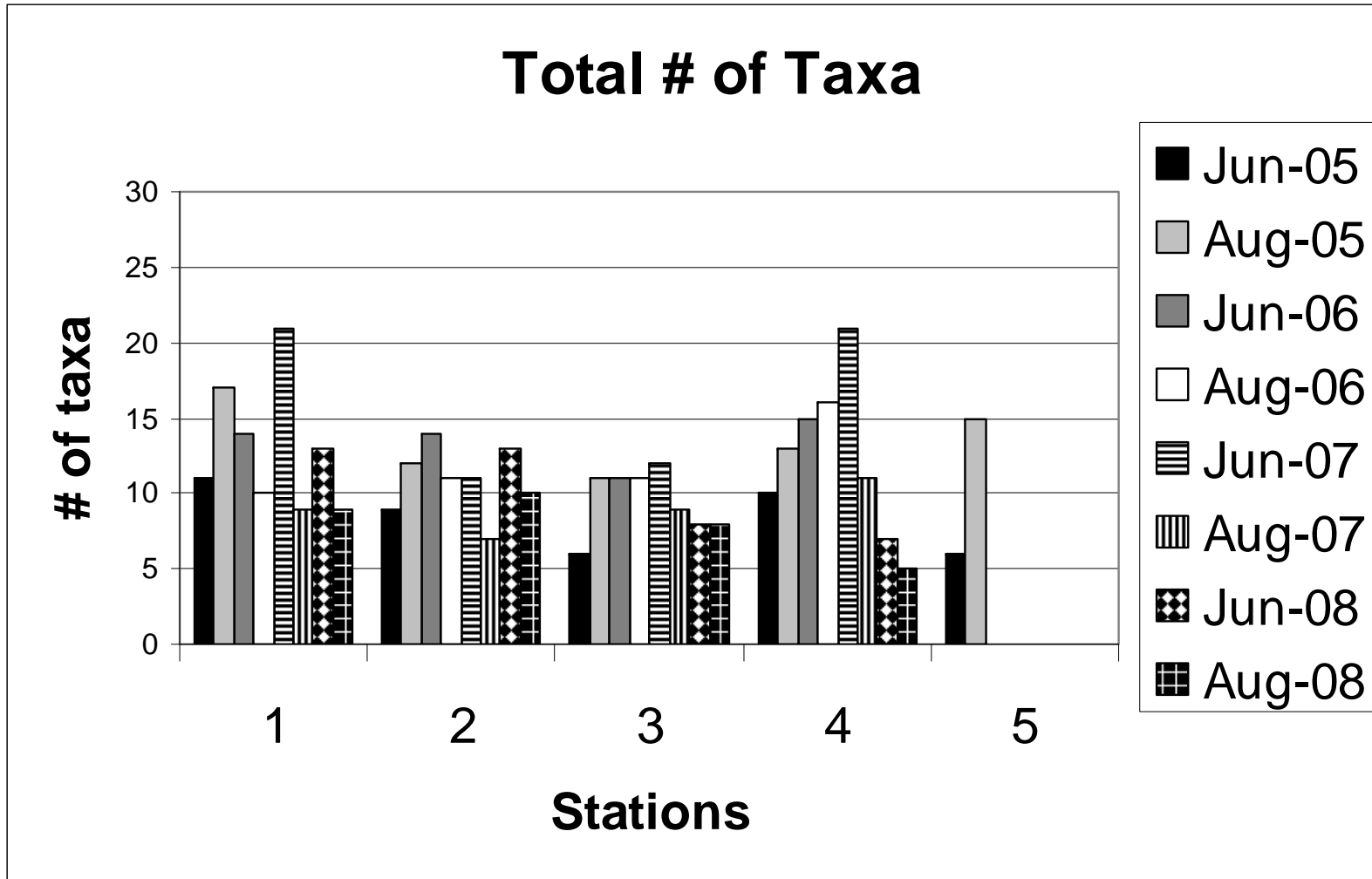


Figure 5. 2005, 2006, 2007 and 2008 benthic macroinvertebrate taxa abundance over space and time in the Mill River, downstream of Lake Whitney. Macroinvertebrate abundance is based on two timed, two minute D frame net samples.

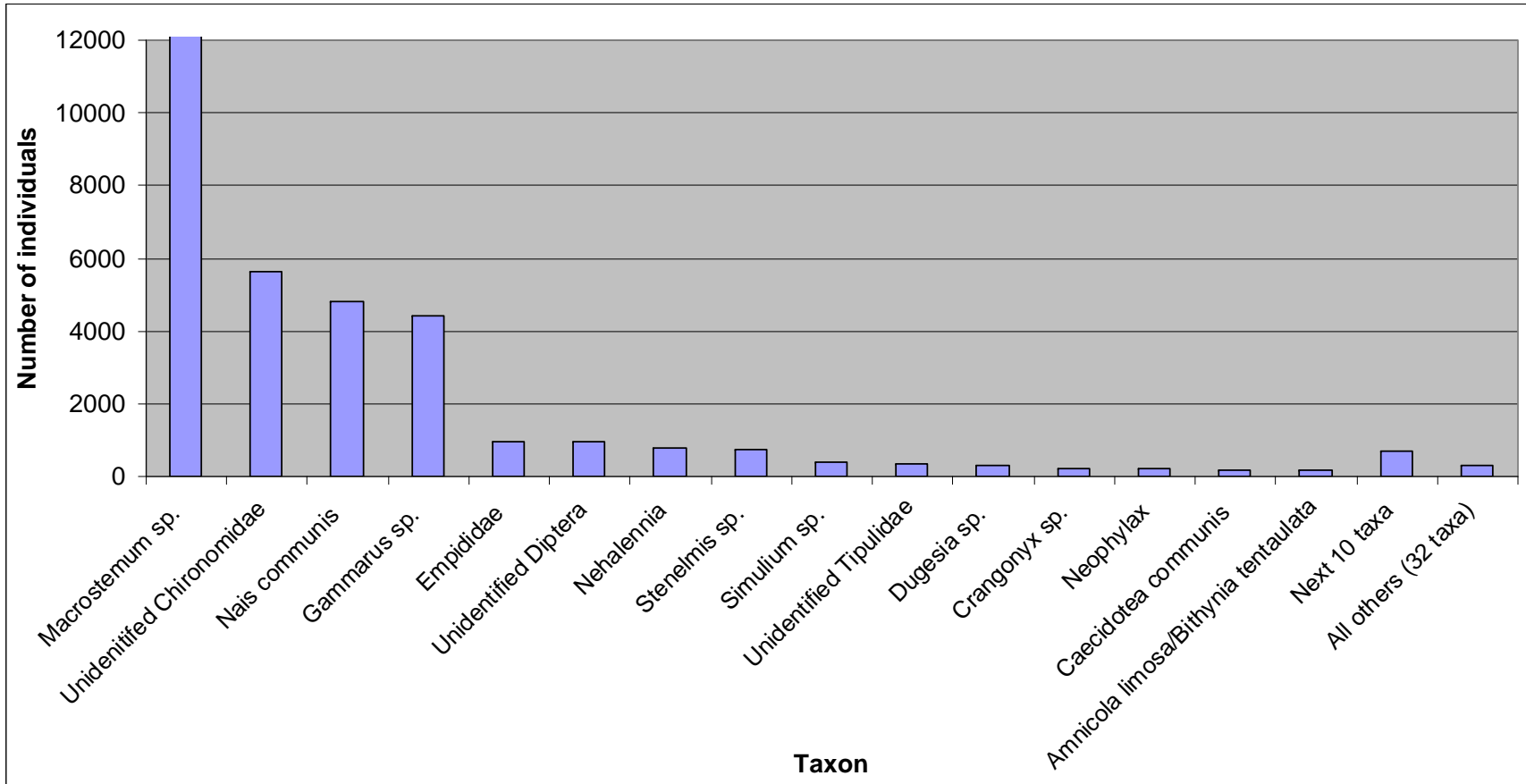


Figure 6. Pooled invertebrate abundance data for 2005, 2006, 2007 and 2008 at all stations in the Mill River, downstream of Lake Whitney. The 15 most abundant invertebrate taxa are graphed, after which the next 10 most abundant are grouped and the remaining individuals are grouped (32 taxa).

Station 1- Feeding Group Abundance

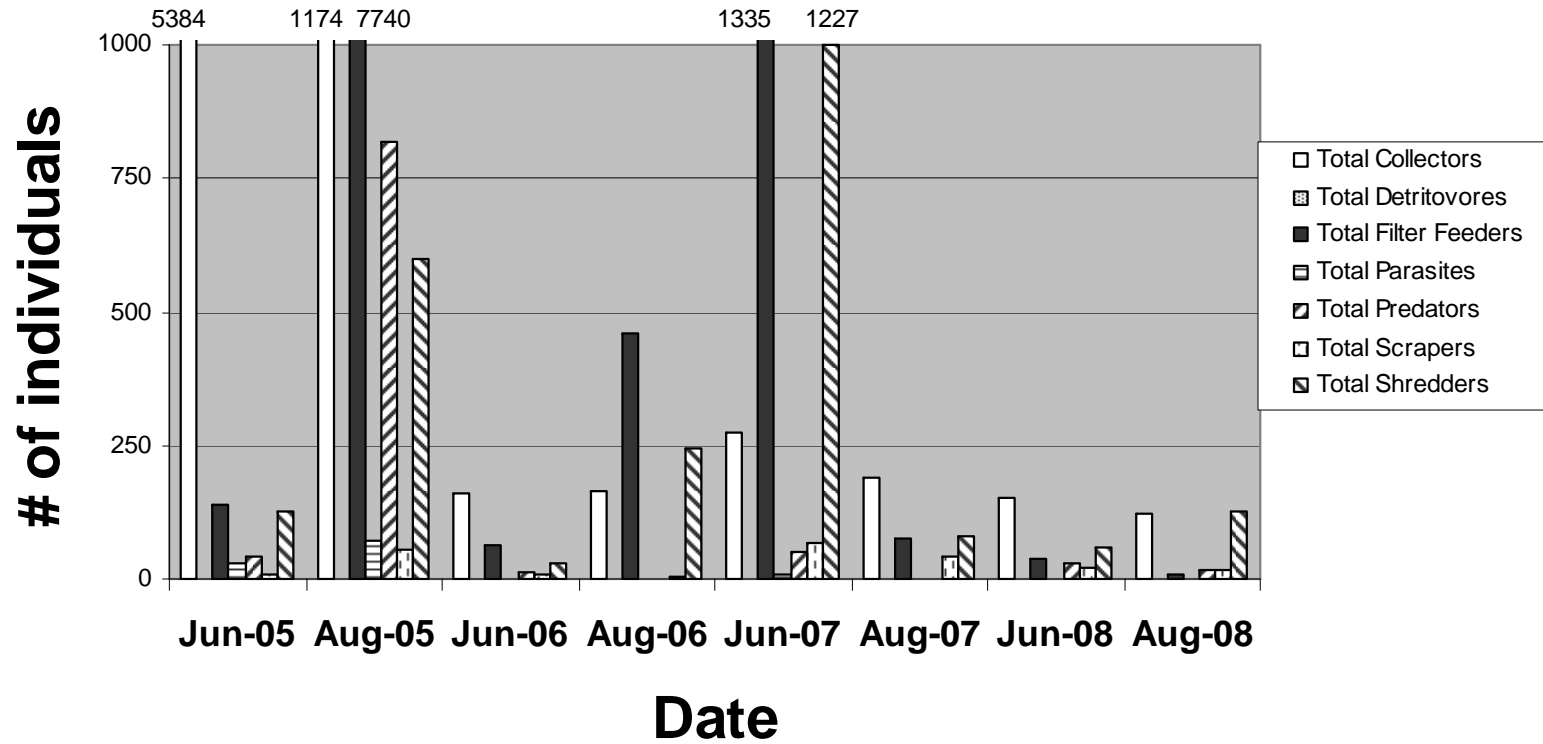


Figure 7. Feeding group presence at Station 1 in 2005, 2006, 2007 and 2008.

Station 2- Feeding Group Abundance

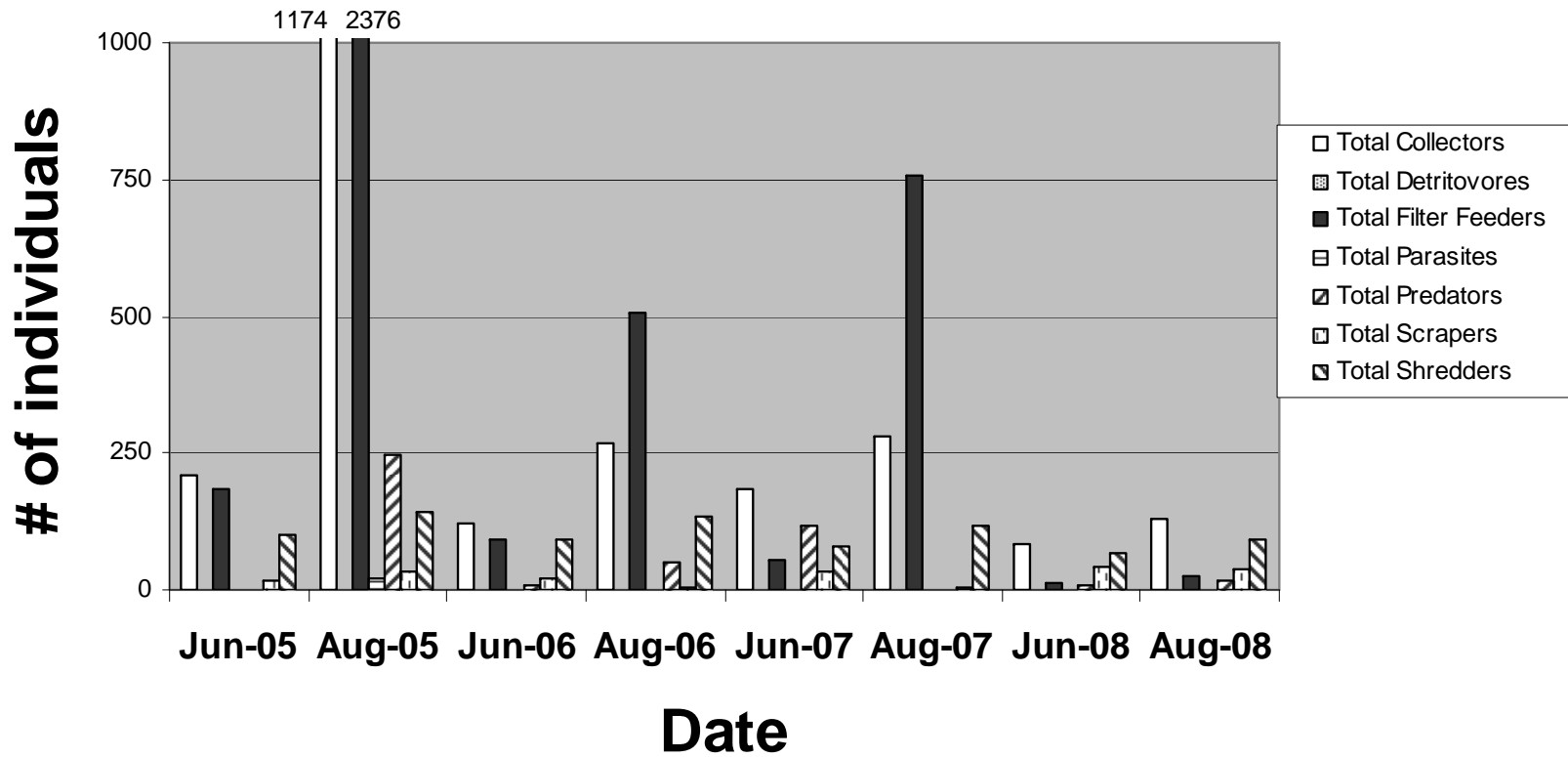


Figure 8. Feeding group presence at Station 2 in 2005, 2006, 2007 and 2008.

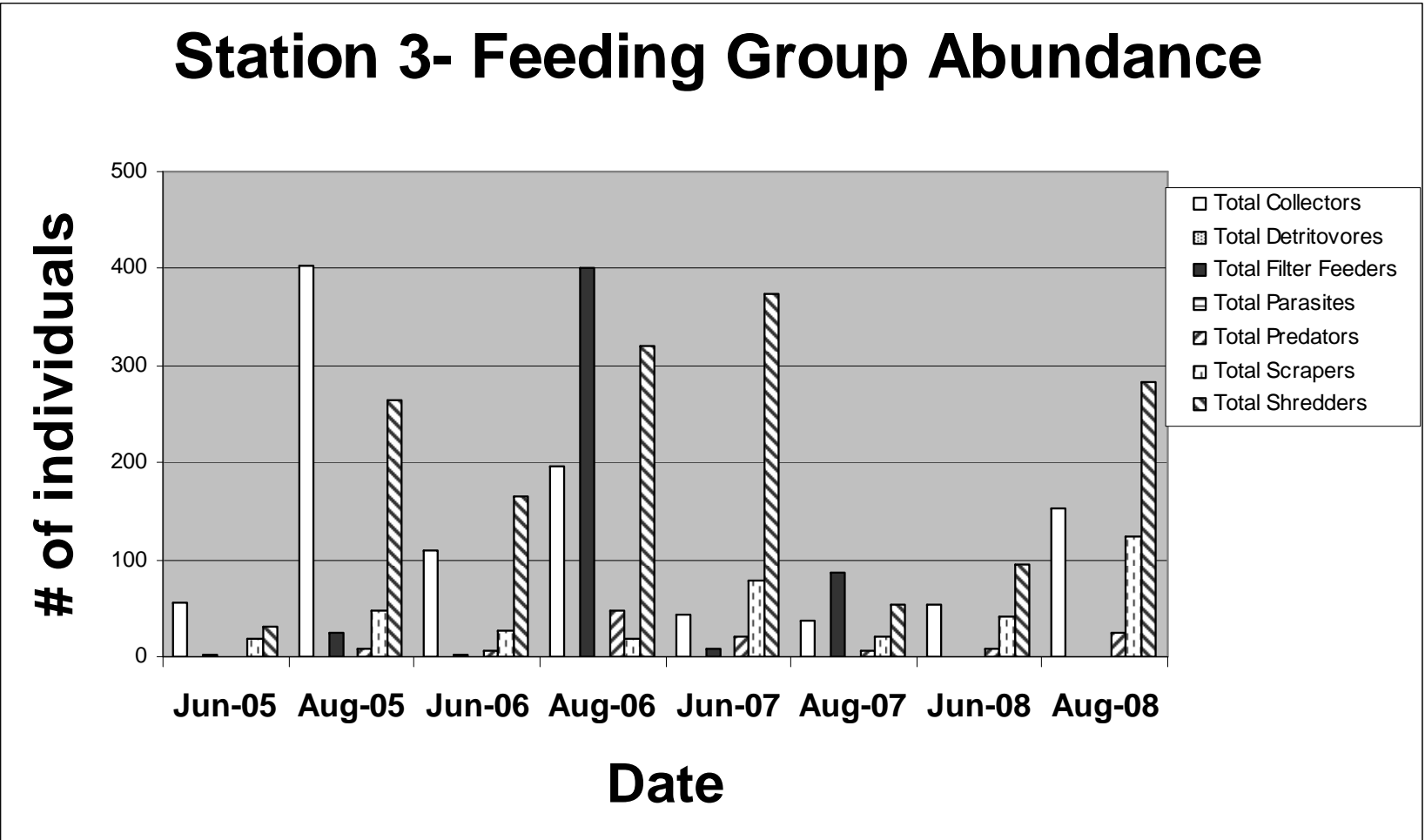


Figure 9. Feeding group presence at Station 3 in 2005, 2006, 2007 and 2008.

Station 4- Feeding Group Abundance

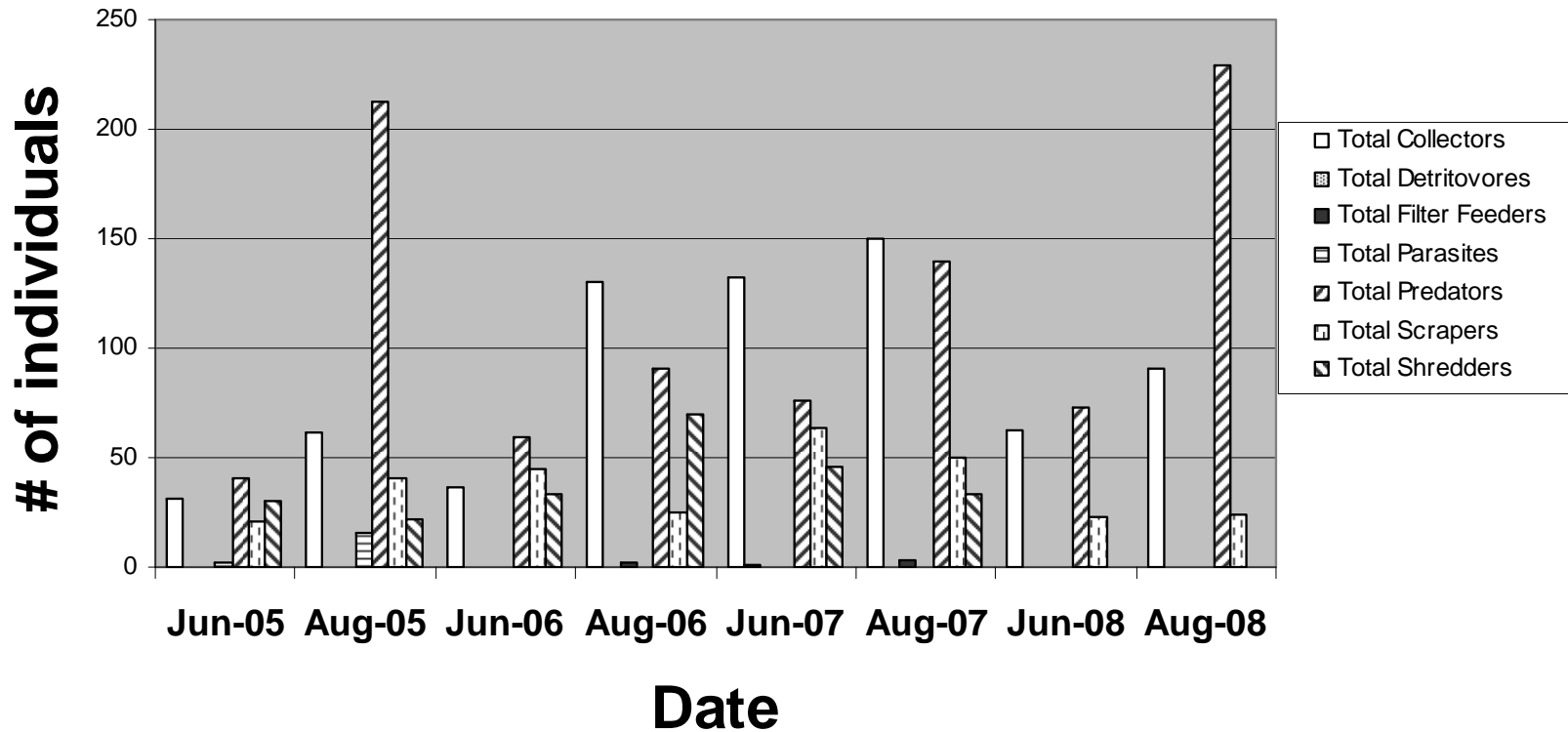


Figure 10. Feeding group presence at Station 4 in 2005, 2006, 2007 and 2008.

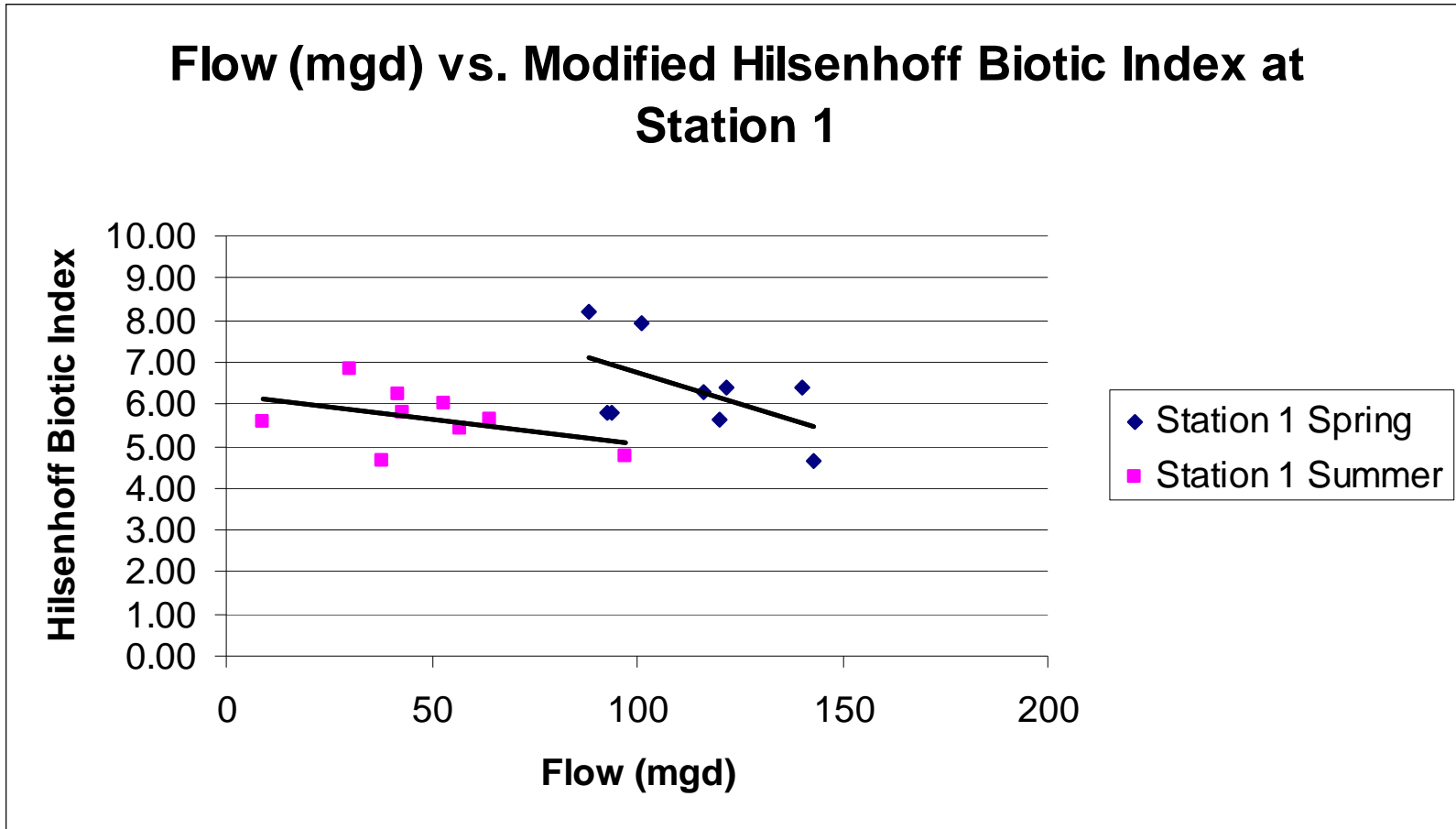


Figure 11. A graph of HBI values vs. average flow for 10 weeks prior to macroinvertebrate sampling for 2000-2008 at station 1. Flow values are based on water flow over the dam, downstream releases and blowoff.

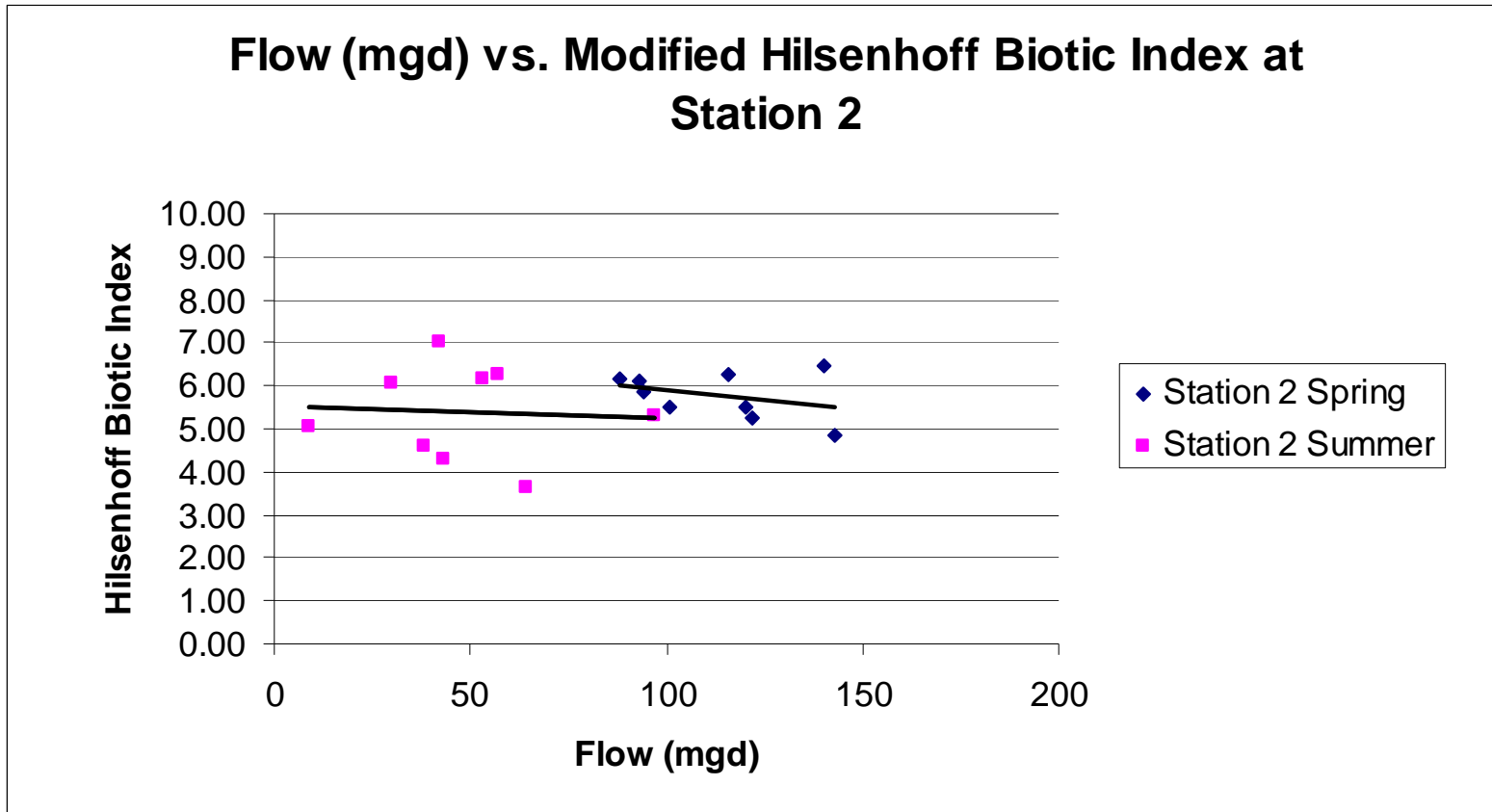


Figure 12. A graph of HBI values vs. average flow for 10 weeks prior to macroinvertebrate sampling for 2000-2008 at station 2. Flow values are based on water flow over the dam, downstream releases and blowoff.

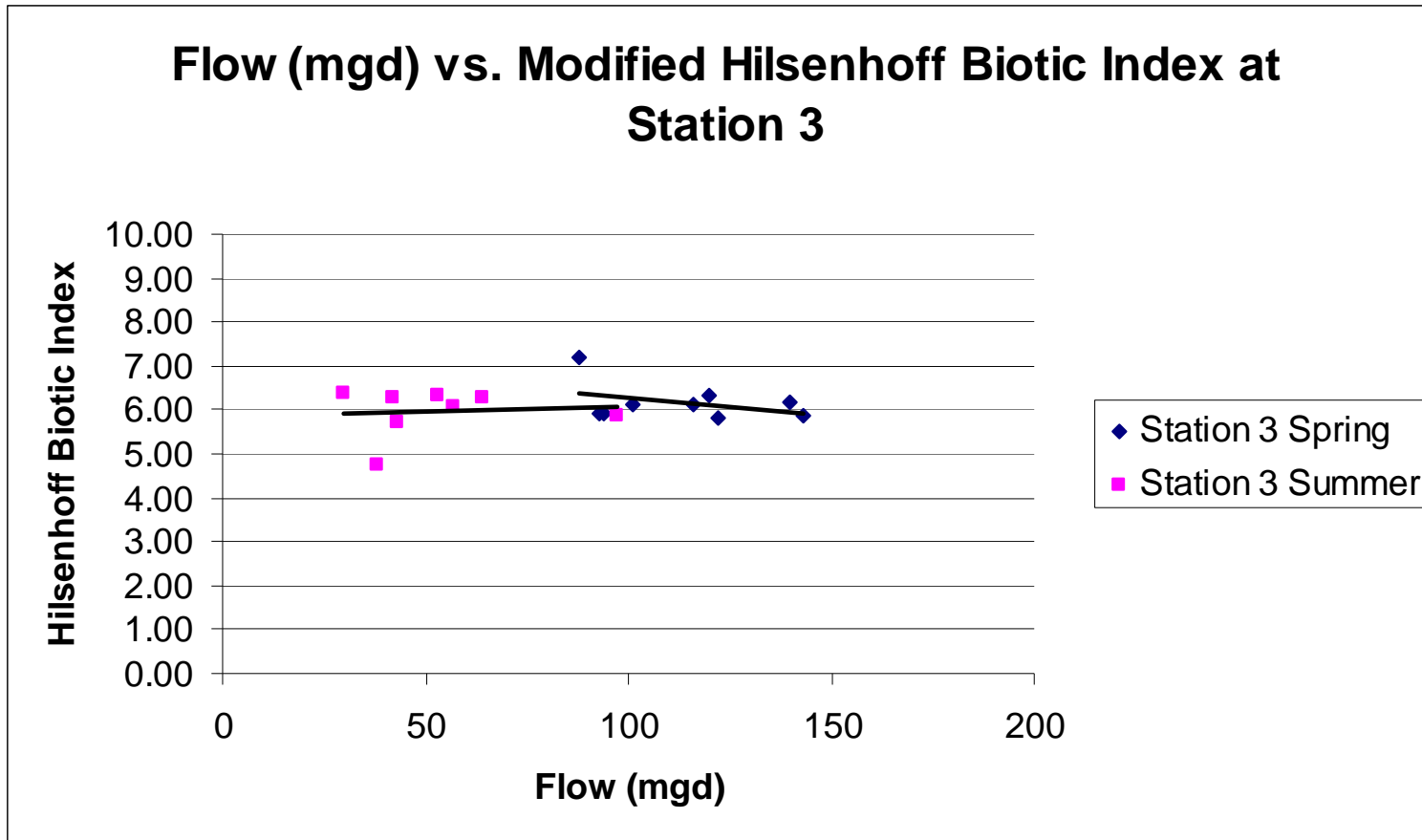


Figure 13. A graph of HBI values vs. average flow for 10 weeks prior to macroinvertebrate sampling for 2000-2008 at station 3. Flow values are based on water flow over the dam, downstream releases and blowoff.

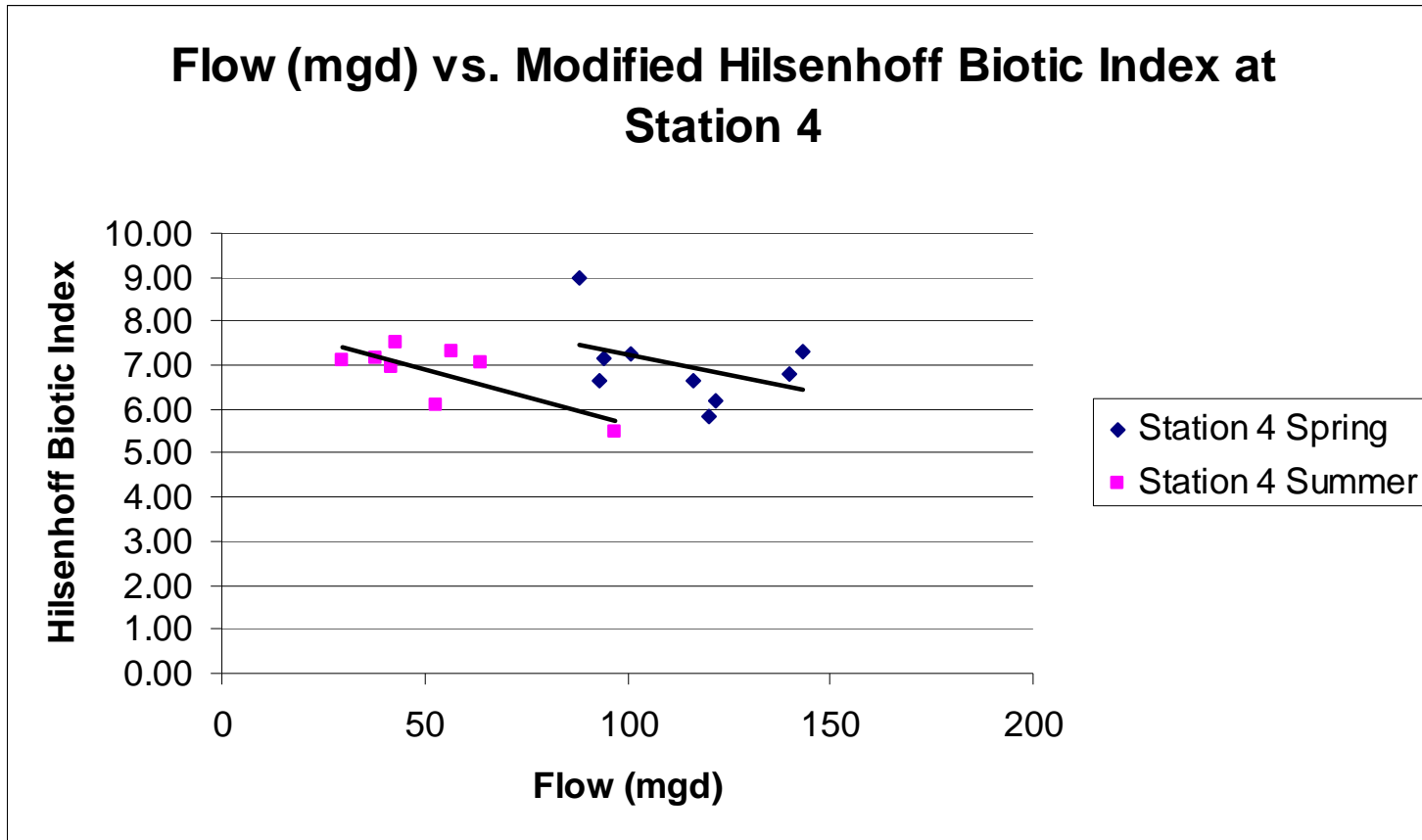


Figure 14. A graph of HBI values vs. average flow for 10 weeks prior to macroinvertebrate sampling for 2000-2008 at station 4. Flow values are based on water flow over the dam, downstream releases and blowoff.

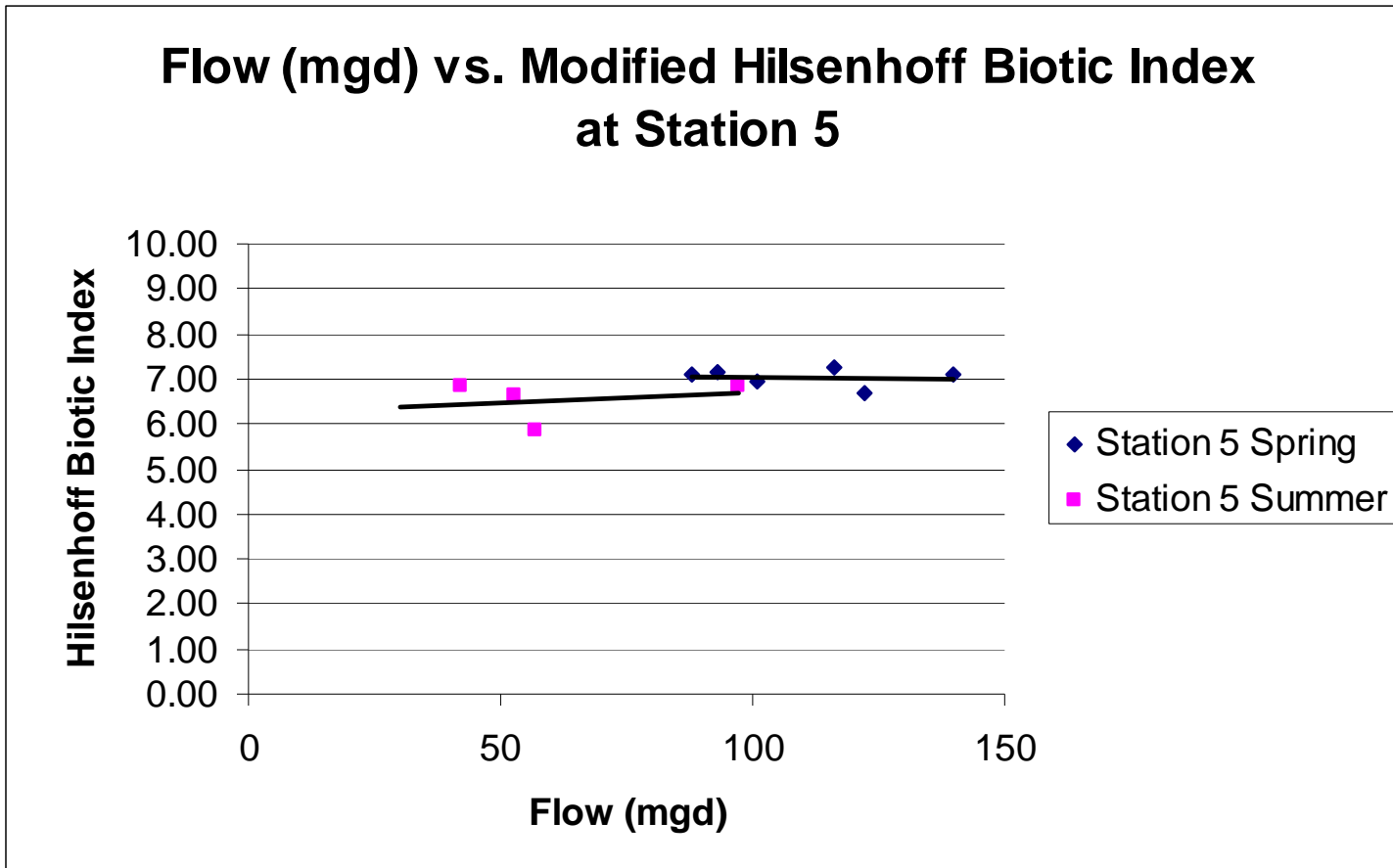


Figure 15. A graph of HBI values vs. average flow for 10 weeks prior to macroinvertebrate sampling for 2000-2005 at station 5. Flow values are based on water flow over the dam, downstream releases and blowoff.

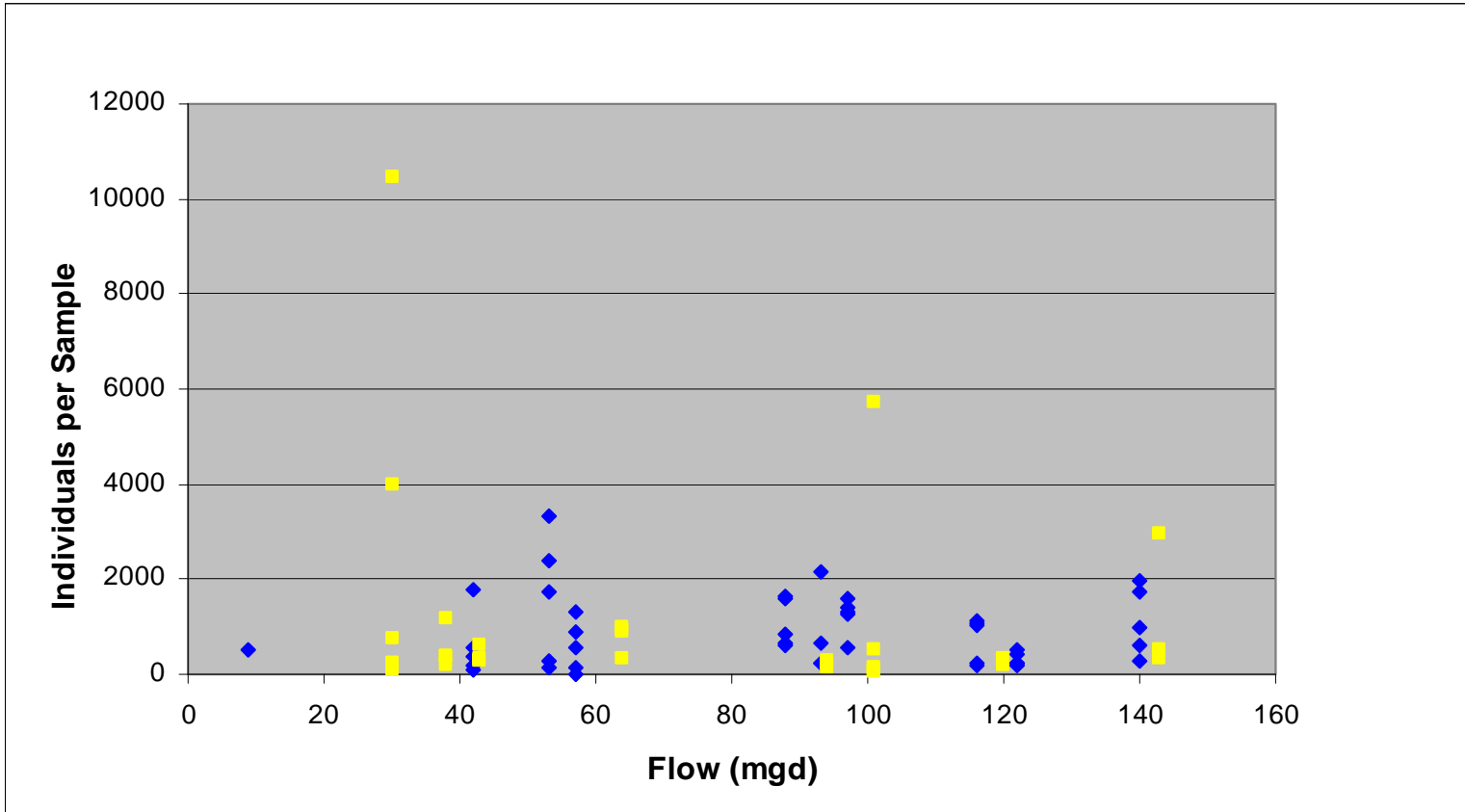


Figure 16. A graph of average flow for 10 weeks prior to macroinvertebrate sampling vs. total number of individuals for 2000-2008 invertebrate data. Pre-operation data is blue and post-operation data is yellow. Flow values are based on water flow over the dam, downstream releases and blowoff.

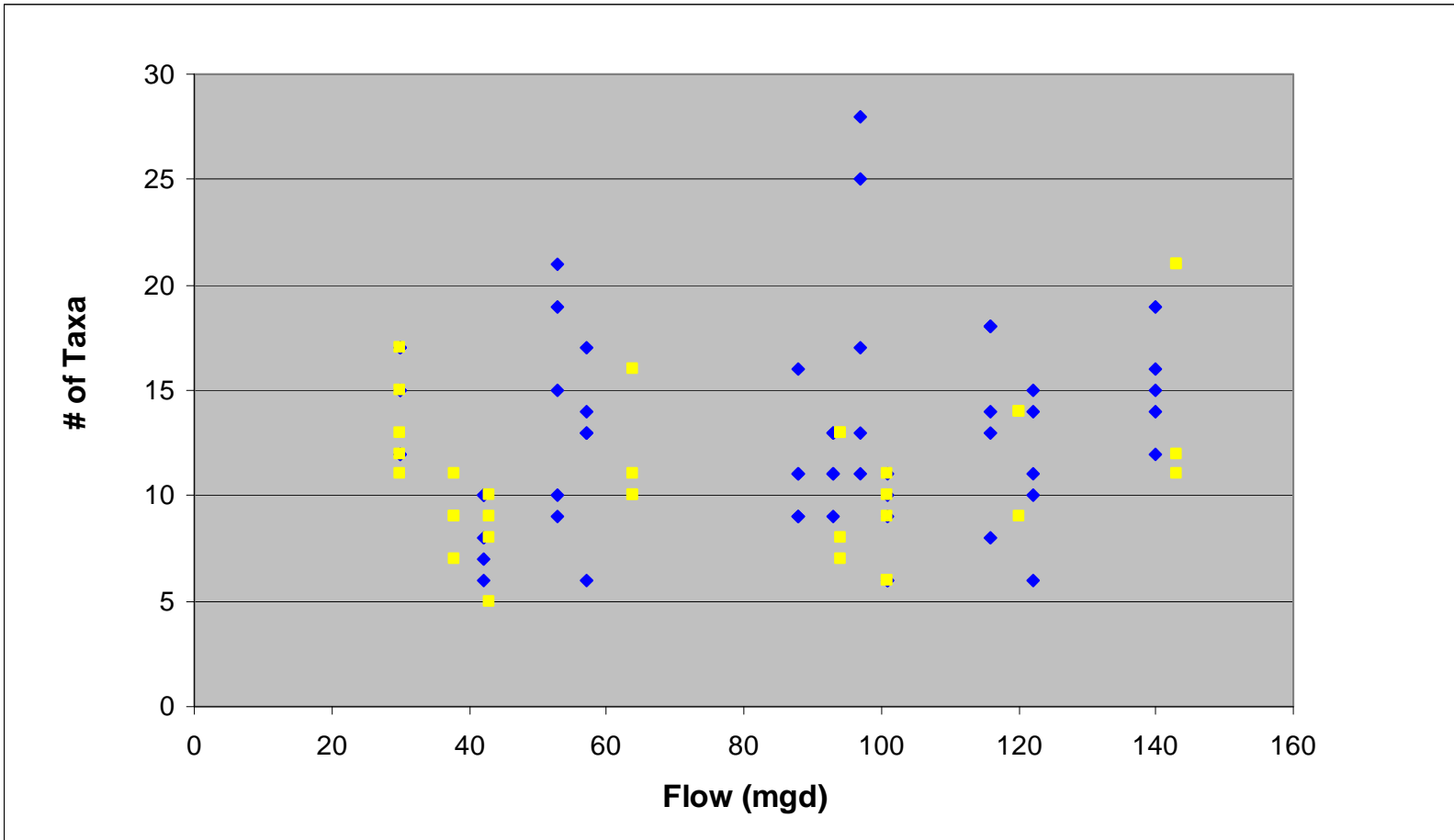


Figure 17. A graph of average flow for 10 weeks prior to macroinvertebrate sampling vs. number of taxa for 2000-2008 invertebrate data. Pre-operation data is blue and post-operation data is yellow. Flow values are based on water flow over the dam, downstream releases and blowoff.

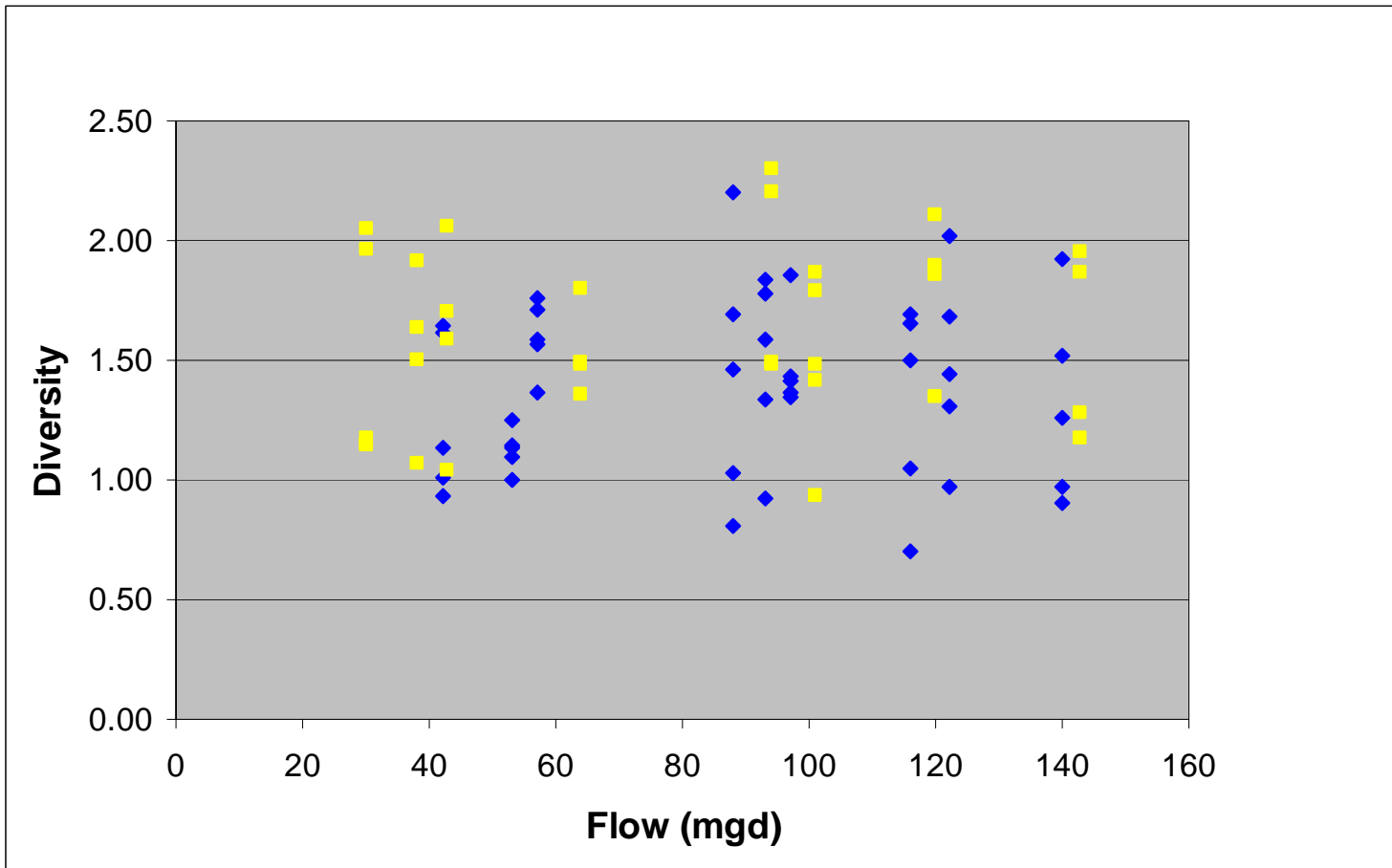


Figure 18. A graph of average flow for 10 weeks prior to macroinvertebrate sampling vs. diversity for 2000-2008 invertebrate data. Pre-operation data is blue and post-operation data is yellow. Flow values are based on water flow over the dam, downstream releases and blowoff.

Table 3. Tabular results of the Modified Hilsenhoff Biotic Index values for 2000-2008 at each station and the corresponding flows. Flow data for September 2004 are not available due to the Lake Whitney drawdown for maintenance.

	2000		2001		2002		2003		2004		2005		2006		2007		2008	
	June	August	June	August	June	August	June	August	June	September	June	August	June	August	June	August	June	August
Station 1	6.3	6.0	6.4	5.4	8.2	6.3	6.4	4.8	5.8	6.8	7.9	5.6	5.7	4.6	4.6	5.8	5.8	5.6
Station 2	6.3	6.2	5.2	6.3	6.2	7.0	6.4	5.3	6.1	6.0	5.5	3.7	5.5	4.6	4.9	4.3	5.8	5.1
Station 3	6.1	6.3	5.8	6.1	7.2	6.3	6.2	5.9	5.9	5.6	6.1	6.4	6.3	6.3	5.9	4.7	5.9	5.7
Station 4	6.7	6.1	6.2	7.3	9.0	6.9	6.8	5.5	6.7	6.8	7.2	7.1	5.8	7.1	7.3	7.2	7.2	7.5
Station 5	7.2	6.6	6.7	5.9	7.1	6.8	7.1	6.8	7.1	7.4	6.9	6.9	*	*	*	*	*	*
Flow (mgd)	116	53	122	57	88	42	140	97	93		101	30	120	64	143	38	94	43

* Station 5 was eliminated from the sampling program in 2006

Table 4. Modified HBI values with suggested water quality designation and degree of organic pollution. Table taken from Mandeville 2002.

Biotic Index	Water Quality	Degree of Organic Pollution
0.00-3.50	Excellent	No apparent organic pollution
3.51-4.50	Very Good	Possible slight organic pollution
4.51-5.50	Good	Some organic pollution
5.51-6.50	Fair	Fairly significant organic pollution
6.51-7.50	Fairly Poor	Significant organic pollution
7.51-8.50	Poor	Very significant organic pollution
8.51-10.00	Very Poor	Severe organic pollution

2008 Lower Mill River Chironomid Taxonomic Study

Analysis of Mill River chironomids from all 2008 samples collected at stations 1-4 during June and August was conducted to assess variability in responses to hydrologic changes among subclassifications of chironomids beyond family level. Previous identification to just the family level was consistent with the methods used for other invertebrates. Further identification of chironomids involves additional sample preparation and examination at higher magnification, which was performed on a subset of previous samples in 2005 to assess potential richness increases. On the advice of a member of the Whitney Environmental Study Team that provides oversight for the overall environmental monitoring program, the SCCRWA agreed to expand sample analysis going forward to include this more detailed assessment of chironomids, although there is no pre-operational data to which these new data can be compared. Identifications followed Epler (2001) a recent and standard reference for this group, with consideration of Simpson and Bode (1980), an older but more regionally appropriate text.

The results presented in Table 4 demonstrate moderate taxonomic richness and fairly consistent composition between stations and dates. There were a total of ten (10) taxa identified in 2008, representing three sub-families of the Chironomidae. Four species have occurred in a majority of samples from 2006-2008. The dominant taxon varied slightly between stations and sampling dates. The two common chironomids in the June 2008 samples, in order of abundance were *Dicrotendipes neomodestus* and *Polypedilum flavum*. Both taxa are common, and were also the dominant species encountered during the 2006 and 2007 analyses. In August, these taxa were present but there was a slight shift in abundance. In August, the most common taxon present was *Polypedilum flavum*, followed by *Cricotopus trifascia*, and *Eukiefferiella tirolensis*. *Dicrotendipes neomodestus* was present in the August samples but ranked fourth in abundance. Remaining chironomid taxa were found at low densities in just a few samples. Compared to 2006 and 2007, chironomid species richness (8 taxa vs. 14 taxa) in 2008 was within the range of the two previous years. The total number of individuals decreased between 2006, 2007 and 2008, although the overall number of total macroinvertebrates decreased during the same period. It is important to note however, that dominant species are similar over all three years.

The ecological indications of virtually all encountered species were of minimal water quality preference (found in a wide range of chemical conditions), high tolerance for elevated nutrients and organic matter (eutrophic conditions), and wide tolerance of current speed with a general preference for moderate to high velocities. The ecological indications of the chironomid species present in the Mill River downstream of the Lake Whitney dam are entirely consistent with observed conditions.

Table 5. Tabular results of the 2006, 2007 and 2008 Chironomid analysis.

Taxon	1-Jun-06				17-Aug-06				19-Jun-07				15-Aug-07				6-Jun-08				19-Aug-08			
	1	2	3	4	1	2	3	4	1	2	3	4	1	2	3	4	1	2	3	4	1	2	3	4
Chironominae																								
Chironomini																								
Chironomus riparius					2																			
Dicrotendipes neomodestus				1	28	80	15	18	42	1	11		12	5	6	5		3	21	1	10	1		
Glyptotendipes lobiferus			5	2	3	6	10	14		3			8	10			1	4	1	1		3		6
Paratendipes albimanus					6	8																		
Polypedilum flavum	30	40	55	8	52	48	24	30	8	11	1		6	31	5	7	9	6		3		5	13	5
Polypedilum braseniae				1																		1		
Polypedilum sp.							1																	
Tanytarsini																								
Paratanytarsus sp.							1																	
Rheotanytarsus exiguus group				3		3	4		13						5		1							
Orthoclaadiinae																								
Cardiocladius obscurum												12	2	2	2			4		2	2			
Cricotopus trifascia	5	3	17	2	10		30	22				23	10	14	9	7	3	1		3	9	5	3	4
Cricotopus intersectus		5		2	4	6	12		2	3			1	4			6						4	2
Cricotopus tibialis		1																						
Cricotopus sylvestris		3	1														1							
Eukiefferiella tirolensis	8																				7	5	3	1
Tanypodinae																								
Procladius sp.											2								1					

DISCUSSION

Hydrologic conditions in 2008 varied between sampling events, but flows did not appear to have a negative impact on the benthic ecology of the river. Channel width and depth at each station were comparable to the range of values measured previously but both decreased between June and August as expected with decreased summer flows. Although August sampling occurred after a dry period, there was still an adequate amount of suitable benthic habitat available for macroinvertebrates and other aquatic organisms. During the August 2008 sampling event, water color at stations 1-3 was light green and was likely due to influences from Lake Whitney. In 2007, ENSR observed moderate numbers of adult blue crabs at station 1. In 2008, no blue crabs of any size were observed or captured.

Differences in macroinvertebrate taxonomic composition between the upstream (stations 1 through 3) and downstream station (station 4) may be ascribed mostly to differences in physical habitat and salinity exposure. However, all salinity measurements by the SCCRWA in summer 2008 at Station 4 were within freshwater values. The August 2008 sampling deviated from observations in previous years as station 4 invertebrate abundance was higher than stations 1 and 2, and may be related to reduced exposure to salinity in 2008 compared to previous years. Freshwater invertebrate tolerance to salinity is not well known, but some of the taxa found in the lower Mill River during previous years (e.g., scuds, damselflies, chironomid midges, beetles, and pulmonate snails) are found in relatively high numbers in moderately saline lakes (Colburn 1988; Alcocer et al. 1998).

The strikingly high invertebrate abundance at Stations 1 and 2 in 2005 was followed by a return to more typical densities in 2006. Initial recolonization of new substrates at station 1 in 2005, along with favorable flows and velocities after a period of diversion at Stations 1 and 2, likely resulted in the observed increases in abundance. It is possible however, that these inflated values could not be sustained long-term and declined in 2006, especially with elevated water velocities as experienced during multiple large storms and elevated summer downstream flows associated with a drawdown of the reservoir. High numbers of invertebrates were again present in the June 2007 sample at station 1. As suggested in 2006, decreased abundance in 2008 could be related to the inability of the habitat to support the inflated number of organisms observed in 2007.

2008 macroinvertebrate levels at station 1 did not approach levels observed in 2005 or 2007. Station 1 experienced a nearly 10-fold decrease in macroinvertebrate abundance between June 2007 and June 2008. Increased abundance in June 2007 was influenced by large numbers of 2 taxa (*Macrostemum* and *Gammurus*). In fact, the increased abundance of macroinvertebrates observed from 2005-2007 is a result of high numbers of these two taxa. 2008 abundance of these taxa was greatly reduced compared to previous years. It's important to note the benthic macroinvertebrates exhibit patchy distribution, and although efforts are made during sampling to

reduce the effects of this type of distribution, it is possible that high numbers of these taxa were present in some portion of each sampling location but were not collected.

No clear patterns are apparent in the 2008 feeding group analysis. Changes between June and August samplings varied between stations and may be related to changes in habitat or water quality, although no significant differences were noted on the days of sampling. Feeding groups present at one station during June were also present in August, but no feeding groups that were not present in June appeared in August. Feeding group distribution was similar between stations but relative abundance of each feeding group was not necessarily similar to nearby stations. Major shifts in feeding groups tend to be related to shifts in individual species abundance. For example, *Macrostemum* sp., a filter feeding caddisfly, has been abundant in most samples since 2003, but were reduced in 2008, which is reflected in the decreased number of filter feeders present.

In general, the macroinvertebrate assemblages observed in the Mill River were indicative of intermediate stream community health. The taxa collected at the four stations located along the Mill River may be commonly found in a range of environments (e.g., worms, scuds, prosobranch snails, caddisflies, mayflies). HBI values at stations 1-3 were within the good to fair categories for most years while stations 4 and 5 were within the fairly poor category (Table 3). Most taxa found were typical of urban freshwater habitats (Walsh et al. 2001), where water quality impacts are common. Midges (Diptera, Chironomidae) which were common invertebrates, can be found in a variety of freshwater habitats (Wetzel 2001c), but their dominance in a community is often regarded as a sign of degraded conditions. In 2008, there was a decrease in *Macrostemum* sp., a species that is less tolerant of pollution. However, members of the family Perlidae were collected for the first time in 2008, and these species are less tolerant of pollution than *Macrostemum*. Even with changes in species abundance between years, the HBI values in 2006, 2007 and 2008 were similar.

This study represents the fourth year of post-operational macroinvertebrate data related to the withdrawal of water in Lake Whitney. As such, although we have attempted to make comparisons, not enough data have been collected to facilitate longer term comparisons among sites or within sites over time as they relate to the activation of the water treatment facility. Initial impressions from these data should be tempered with the larger data set that will be generated over the course of the planned study. Comparisons of flow and pre and post-operational data available thus far do not suggest any impacts of the water treatment facility going online.

As noted in the summary report for the 2000-2004 pre-operational monitoring program, changes in the invertebrate community over time may be a consequence of many environmental factors, including the desiccation of the stream during the dry summer months, changes in water quality, altered food abundance and quality, and predation effects. Flow is only one factor, and is likely to have more indirect effects at low levels. Variability in flow, inducing instability, may also be a

potent factor in structuring the benthic macroinvertebrate community of the lower Mill River, and is linked to water quality issues (including dilution of contaminants from upstream and salinity from downstream), altered physical habitat, and available food resources.

Reduced flow may decrease invertebrate density and diversity (Gørtz 1998; Brunke et al. 2001), but flow interacts closely with the physical structure of the habitat. Streams with relatively low flow but a high degree of habitat heterogeneity (coarse detritus, rocks, submerged vegetation) may still support high invertebrate density, taxonomic richness and diversity (Brunke et al. 2001). Increased vegetation cover may be expected at lower flow regimes, thus counterbalancing (at least in part) the potentially negative effects of decreased flow by increasing substrate heterogeneity. Relatively rapid response of invertebrate communities suggests that recovery will occur within months after a drought period. In the Mill River, macroinvertebrate density tends to increase slightly with decreasing flow and is related to decreased stream width and depth.

Effects of increased salinity on the lower Mill River invertebrate assemblages are difficult to predict, but would seem likely to be more severe than minor changes in flow. Reduced freshwater flow could increase salinity effects. Most of the taxa found in this survey may withstand small increases in salinity, with invertebrate communities shaped more by physical habitat characteristics than those fluctuations in salinity (Alcocer et al. 1998). However, effects of possible tide-related bursts in salinity, exacerbated by lower flow or removal of tide gates, could shift the community to a taxa-poor, low-diversity assemblage dominated by high salinity tolerant taxa (Wolfram et al. 1999). The current community at station 4, where salinity exposure is periodically elevated, already exhibits this condition. Due to its higher elevation, the upstream portion of the lower Mill River (stations 1 through 3) is isolated from saltwater intrusion and thus appears unlikely to be affected by tide-driven salinity bursts.

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APPENDIX A

2005-2008 Benthic Macroinvertebrate Data

Class	Order	Family	Genus/Species	Feeding Group	1-Jun-06				1-Aug-06					
					Stations				Stations					
					1	2	3	4	1	2	3	4		
Annelida	Hirudinea	Glossiphoniidae	Glossiphonia complanata	Parasite										
Annelida	Hirudinea	Glossiphoniidae	Placobdella sp.	Parasite										
Annelida	Hirudinea		Hirudina	Parasite										
Annelida	Oligochaeta	Lumbriculidae	Unidentified Lumbriculidae	Collector										
Annelida	Oligochaeta	Naididae	Nais communis	Collector	14	2			5					
Annelida	Oligochaeta	Oligochaeta	Unidentified Oligochaeta	Collector				4						
Annelida	Oligochaeta	Tubificidae	Unidentified Tubificidae	Collector										
Annelida	Oligochaeta	Tubificidae	Limnodrilus hoffmeisteri	Collector										
Annelida	Polychaeta	Ampheriidae	Unidentified Ampheriidae	Detritivore										
Annelida	Polychaeta	Capitellidae	Heteromastus filiformis	Detritivore										
Annelida	Polychaeta	Spionidae	Marenzelleria viridis	Filter Feeder										
Annelida	Polychaeta	Spionidae	Polydora sp.	Detritivore										
Arachnida	Trombidiformes	Lebertidae	Lebertia sp.	Predator									5	
Arachnoidea	Hydracarina	Arrenuridae	Unidentified Arrenuridae	Parasite										
Bivalvia	Veneorida	Pisidiidae	Pisidium sp.	Filter Feeder										2
Branchiopoda	Cladocera		cladocera	Collector										
Crustacea	Amphipoda	Corophidae	Corophium sp. (juvenile)	Filter Feeder										
Crustacea	Amphipoda	Crangonyctidae	Crangonyx sp.	Shredder	4	5	6	1						
Crustacea	Amphipoda	Gammaridae	Gammarus sp.	Shredder	10	83	161	31	241	117	316	70		
Crustacea	Cumacea	Nannastacidae	Almyracuma proximoculi	Shredder										
Crustacea	Decapoda	Palaemonidae	Palaemonetes vulgaris	Shredder										
Crustacea	Decapoda	Palaemonidae	Palaemonetes paludosus	Shredder										
Crustacea	Decapoda	Portunidae	Carcinus maenus	Shredder										
Crustacea	Isopoda	Aseilidae	Caecidotea communis	Collector	34	36	0	1	4	0	4	6		
Crustacea	Isopoda	Aseilidae	Lirceus/Acellus sp. (communis)	Shredder										
Hydrozoa	Hydrida	Hydridae	Hydra sp.	Predator										
Insecta	Coleoptera	Brachyceridae	Brachycerus sp.	Collector										8
Insecta	Coleoptera	Chrysomelidae	Donacia	Shredder										
Insecta	Coleoptera	Coleoptera	Unidentified Coleoptera	Predator										
Insecta	Coleoptera	Curculionidae	Unidentified Curculionidae	Shredder										
Insecta	Coleoptera	Dryopidae	Helichus sp.	Predator										
Insecta	Coleoptera	Elmidae	Stenelmis sp.	Scraper	8	2	20	44	5	4	8	18		
Insecta	Coleoptera	Halipidae	Peltodytes	Shredder										
Insecta	Coleoptera	Hydrophilidae	Berosus sp.	Predator	7	1	7		1					
Insecta	Coleoptera	Psephenidae	Unidentified Psephenidae	Predator										
Insecta	Diptera	Atrichopogon	Atrichopogon	Predator										
Insecta	Diptera	Ceratopogonidae	Probezzia	Predator				1						
Insecta	Diptera	Ceratopogonidae	Unidentified Ceratopogonidae	Predator										
Insecta	Diptera	Chironomidae	Unidentified Chironomidae	Collector	61	66	86	24	134	216	142	107		
Insecta	Diptera	Diptera	Unidentified Diptera	Collector	50	15	24	5	22	51	30	4		
Insecta	Diptera	Empididae	Empididae	Predator						49	42			
Insecta	Diptera	Empididae	Hemerodromia sp.	Filter Feeder										
Insecta	Diptera	Simuliidae	Simulium sp.	Filter Feeder	6	1	1		44	37				
Insecta	Diptera	Tabanidae	tabanidae	Predator										
Insecta	Diptera	Tachinidae	Ceracia	Parasite										
Insecta	Diptera	Tipulidae	Unidentified Tipulidae	Shredder	3	3	0	0	4	17	4			
Insecta	Ephemeroptera	Baetidae	Baetis sp.	Collector				1						2
Insecta	Ephemeroptera	Caenidae	Caenis sp.	Collector										
Insecta	Ephemeroptera	Ephemerellidae	Unidentified Ephemerellidae	Collector										
Insecta	Ephemeroptera	Heptageniidae	Stenonema sp.	Scraper										
Insecta	Ephemeroptera	Oligoneuridae	Isonychia sp.	Collector										
Insecta	Hemiptera	Aphididae	aphididae	Predator	1						1			
Insecta	Hemiptera	Gelastocoridae	Gelastocoris	Predator										1
Insecta	Hemiptera	Hemiptera	Unidentified Hemiptera	Predator										
Insecta	Heteroptera	Gerridae	Rheumatobates sp.	Predator										1
Insecta	Heteroptera	Gerridae	Unidentified Gerridae	Predator										
Insecta	Heteroptera	Mesovellidae	Mesovella sp.	Predator										
Insecta	Heteroptera	Veliidae	Microvelia	Predator										
Insecta	Neuroptera	Sisyridae	Sisyra sp.	Predator										
Insecta	Odonata	Calopterygidae	Calopteryx spp	Predator										
Insecta	Odonata	Coenagrionidae	Nehalennia	Predator				50						80
Insecta	Odonata	Coenagrionidae	Ischnura/Enallagma sp.	Predator										
Insecta	Odonata	Coenagrionidae	Argia sp.	Predator		2		3						2
Insecta	Odonata	Cordulegastriidae	Epithecica	Predator										
Insecta	Odonata	Cordulidae	Epicordulia	Predator				5						6
Insecta	Odonata	Cordulidae	Somatochora sp.	Predator										
Insecta	Odonata	Cordulidae	Didymops sp.	Predator										
Insecta	Odonata		Anisoptera (juvenile)	Predator										
Insecta	Odonata		Zygoptera fragmenta	Predator										
Insecta	Trichoptera	Brachycentridae	Brachycentrus sp.	Filter Feeder										
Insecta	Trichoptera	Brachycentridae	Micrasema sp.	Filter Feeder										
Insecta	Trichoptera	Glossosomatidae	Glossosoma	Scraper										
Insecta	Trichoptera	Hydropsychidae	Macrostemum sp.	Filter Feeder	56	92	1		416	468	401			
Insecta	Trichoptera	Hydropsychidae	Hydropsyche sp.	Filter Feeder										
Insecta	Trichoptera	Hydropsychidae	Parapsyche sp.	Filter Feeder										
Insecta	Trichoptera	Hydroptilidae	Agraula sp.	Parasite							1	22		2
Insecta	Trichoptera	Hydroptilidae	Oxyethira sp.	Predator										
Insecta	Trichoptera	Hydroptilidae	Orthotrichia sp.	Predator										
Insecta	Trichoptera	Leptoceridae	Ceraclea sp.	Collector										
Insecta	Trichoptera	Leptoceridae	Mystacides sp.	Collector										
Insecta	Trichoptera	Leptoceridae	Trienodes sp.	Shredder										
Insecta	Trichoptera	Limnephilidae	Rossiana sp.	Scraper										
Insecta	Trichoptera	Limnephilidae	Unidentified Limnephilidae	Scraper				1						
Insecta	Trichoptera	Philopotamidae	Chimarra spp	Filter Feeder										
Insecta	Trichoptera	Psychomyiidae	Psychomyia sp.	Collector										
Insecta	Trichoptera	Uenoidae	Neophylax	Shredder	11			1						
Malacostraca	Amphipoda	Hyalellidae	Hyalella azteca	Collector										
Malacostraca	Decapoda	Cambaridae	Orconectes limosus	Shredder										
Malacostraca	Decapoda	Cambaridae	Unidentified Cambaridae	Shredder										
Molluscopoda	Sessilia	Balanidae	Balanus improvisus	Filter Feeder										
Mollusca	Bivalvia	Sphaeriidae	Unidentified Sphaeriidae	Scraper										
Mollusca	Gastropoda	Ancylidae	Ferrissia rivularis	Scraper										
Mollusca	Gastropoda	Gastropoda	Unidentified Gastropoda	Scraper										
Mollusca	Gastropoda	Hydrobiidae	Amnicola limosa/Bithynia tentaculata	Scraper	1	18	7				10			
Mollusca	Gastropoda	Hydrobiidae	Pomatopsis sp.	Scraper										
Mollusca	Gastropoda	Lymnaeidae	Lymnaea columella	Scraper										
Mollusca	Gastropoda	Physidae	Physa sp.	Scraper										5
Mollusca	Gastropoda	Planorbidae	Gyraulus parvus	Scraper										
Mollusca	Gastropoda	Planorbidae	Helisoma sp.	Scraper										2
Mollusca	Gastropoda	Planorbidae	Gyraulus deflectus	Scraper										
Mollusca	Gastropoda	Planorbidae	Gyraulus circumstriatus	Scraper										
Mollusca	Gastropoda	Pleuroceridae	Pleurocera sp.	Scraper										
Mollusca	Gastropoda	Valvatidae	Valvata tricarinata	Scraper										
Nemertea	Nemertea	Nemertea	Unidentified Nemertea	Predator										
Turbellaria	Tricladida	Dugesidae	Dugesia sp.	Predator	3	7					1			

Class	Order	Family	Genus/Species	6-Jun-08				19-Aug-08					
				Stations				Stations					
				1	2	3	4	1	2	3	4		
Annelida	Hirudinea	Glossiphoniidae	Glossiphonia complanata										
Annelida	Hirudinea	Glossiphoniidae	Placobdella sp.										
Annelida	Hirudinea		Hirudinia										
Annelida	Oligochaeta	Lumbriculidae	Unidentified Lumbriculidae										
Annelida	Oligochaeta	Naididae	Nais communis	30	14			17	45	25			
Annelida	Oligochaeta	Oligochaeta	Unidentified Oligochaeta										
Annelida	Oligochaeta	Tubificidae	Unidentified Tubificidae										
Annelida	Oligochaeta	Tubificidae	Limnodrilus hoffmeisteri										
Annelida	Polychaeta	Ampheriidae	Unidentified Ampheriidae										
Annelida	Polychaeta	Capitellidae	Heteromastus filiformis										
Annelida	Polychaeta	Spionidae	Marenzelleria viridis										
Annelida	Polychaeta	Spionidae	Polydora sp.										
Arachnida	Trombidiformes	Lebertiidae	Lebertia sp.										
Arachnoidea	Hydracarina	Arrenuridae	Unidentified Arrenuridae										
Bivalvia	Veneorida	Pisidiidae	Pisidium sp.										
Branchiopoda	Cladocera		cladocera										
Crustacea	Amphipoda	Corophiidae	Corophium sp. (juvenile)										
Crustacea	Amphipoda	Crangonyctidae	Crangonyx sp.										
Crustacea	Amphipoda	Gammaridae	Gammarus sp.	50	54	83		108	62	242			
Crustacea	Cumacea	Nannastacidae	Almyracuma proximoculi										
Crustacea	Decapoda	Palaemonidae	Palaemonetes vulgaris										
Crustacea	Decapoda	Palaemonidae	Palaemonetes paludosus										
Crustacea	Decapoda	Portunidae	Carcinus maenus										
Crustacea	Isopoda	Asellidae	Caecidotea communis	30	11		5						
Crustacea	Isopoda	Asellidae	Lirceus/Acellus sp. (communis)										
Hydrozoa	Hydroida	Hydridae	Hydra sp.										
Insecta	Coleoptera	Brachymeridae	Brachymerus sp.		4					9			
Insecta	Coleoptera	Chrysomelidae	Donacia										
Insecta	Coleoptera	Coleoptera	Unidentified Coleoptera										
Insecta	Coleoptera	Curculionidae	Unidentified Curculionidae										
Insecta	Coleoptera	Dryopidae	Helichus sp.										
Insecta	Coleoptera	Elmidae	Stenelmis sp.	10	29	38	5	8	36	108			
Insecta	Coleoptera	Haliplidae	Pelodytes										
Insecta	Coleoptera	Hydrophilidae	Berosus sp.			4							
Insecta	Coleoptera	Psephenidae	Unidentified Psephenidae										
Insecta	Diptera	Atrichopogon	Atrichopogon										
Insecta	Diptera	Ceratopogonidae	Probezzia										
Insecta	Diptera	Ceratopogonidae	Unidentified Ceratopogonidae										
Insecta	Diptera	Chironomidae	Unidentified Chironomidae	72	44	53	17	68	38	128	44		
Insecta	Diptera	Diptera	Unidentified Diptera	20				17	36				
Insecta	Diptera	Empididae	Empididae	10									
Insecta	Diptera	Empididae	Hemerodromia sp.										
Insecta	Diptera	Simuliidae	Simulium sp.	30	11			8	18				
Insecta	Diptera	Tabanidae	tabanidae										
Insecta	Diptera	Tachinidae	Ceracia										
Insecta	Diptera	Tipulidae	Unidentified Tipulidae	10	4	4				9			
Insecta	Ephemeroptera	Baetidae	Baetis sp.		11								
Insecta	Ephemeroptera	Caenidae	Caenis sp.				41						47
Insecta	Ephemeroptera	Ephemerellidae	Unidentified Ephemerellidae										
Insecta	Ephemeroptera	Heptageniidae	Stenonema sp.										
Insecta	Ephemeroptera	Oligoneuridae	Isurynchia sp.										
Insecta	Hemiptera	Gelastocoridae	Gelastocoris										
Insecta	Hemiptera	Hemiptera	Unidentified Hemiptera										
Insecta	Heteroptera	Gerridae	Rheumatobates sp.										
Insecta	Heteroptera	Gerridae	Unidentified Gerridae										
Insecta	Heteroptera	Mesoveliidae	Mesovella sp.										
Insecta	Heteroptera	Yellidae	Microvelia	10									
Insecta	Neuroptera	Sisyridae	Sisyra sp.										
Insecta	Odonata	Calopterygidae	Calopteryx spp										
Insecta	Odonata	Coenagrionidae	Nehalennia				73						229
Insecta	Odonata	Coenagrionidae	Ischnura/Enallagma sp.										
Insecta	Odonata	Coenagrionidae	Argia sp.			4				8			
Insecta	Odonata	Cordulegastridae	Epithea										
Insecta	Odonata	Cordulidae	Epicordulia										
Insecta	Odonata	Cordulidae	Somatochlora sp.										
Insecta	Odonata	Cordulidae	Didymops sp.										
Insecta	Odonata		Anisoptera (juvenile)										
Insecta	Odonata		zygoptera fragments										
Insecta	Plecoptera	Perlidae	Perlidae sp.	10	7			17	18	17			
Insecta	Trichoptera	Brachycentridae	Brachycentrus sp.										
Insecta	Trichoptera	Brachycentridae	Micrasema sp.										
Insecta	Trichoptera	Glossogomphidae	Glossosoma										
Insecta	Trichoptera	Hydropsychidae	Macroctenium sp.	10						9			
Insecta	Trichoptera	Hydropsychidae	Hydropsyche sp.										
Insecta	Trichoptera	Hydropsychidae	Parapsyche sp.										
Insecta	Trichoptera	Hydroptilidae	Agraylea sp.										
Insecta	Trichoptera	Hydroptilidae	Oxethira sp.										
Insecta	Trichoptera	Hydroptilidae	Orthotrichia sp.										
Insecta	Trichoptera	Leptoceridae	Ceraclea sp.										
Insecta	Trichoptera	Leptoceridae	Mystacides sp.										
Insecta	Trichoptera	Leptoceridae	Trienodes sp.										
Insecta	Trichoptera	Limnephilidae	Rossiana sp.										
Insecta	Trichoptera	Limnephilidae	Unidentified Limnephilidae										
Insecta	Trichoptera	Philopotamidae	Chimarra spp										
Insecta	Trichoptera	Psychomyiidae	Psychomyia sp.										
Insecta	Trichoptera	Uenoidae	Negophylax		11	8		17		42			
Malacostraca	Amphipoda	Hyalellidae	Hyalella azteca										
Malacostraca	Decapoda	Cambaridae	Orconectes limosus										
Malacostraca	Decapoda	Cambaridae	Unidentified Cambaridae										
Madillopoda	Sessilia	Balanidae	Balanus improvisus										
Mollusca	Bivalvia	Sphaeriidae	Unidentified Sphaeriidae										
Mollusca	Gastropoda	Ancylidae	Ferrissia rivularis										
Mollusca	Gastropoda	Gastropoda	Unidentified Gastropoda	10				8					
Mollusca	Gastropoda	Hydrobiidae	Ammicola limosa/Bithynia tentaculata		7					17			
Mollusca	Gastropoda	Hydrobiidae	Pomatopsis sp.										
Mollusca	Gastropoda	Lymnaeidae	Lymnaea columella										
Mollusca	Gastropoda	Physidae	Physa sp.			4	9						12
Mollusca	Gastropoda	Planorbidae	Gyraulus parvus										
Mollusca	Gastropoda	Planorbidae	Helisoma sp.	4			9						12
Mollusca	Gastropoda	Planorbidae	Gyraulus deflectus										
Mollusca	Gastropoda	Planorbidae	Gyraulus circumstriatus										
Mollusca	Gastropoda	Pleuroceridae	Pleurocera sp.										
Mollusca	Gastropoda	Valvatidae	Valvata tricarinata										
Nemertea	Nemertea	Nemertea	Unidentified Nemertea										
Turbellaria	Tricladida	Dugesidae	Dugesia sp.										